

European Journal of Wildlife Research

Determining the subspecies composition of bean goose harvests in Finland using genetic methods --Manuscript Draft--

Manuscript Number:	EJWR-D-16-00127R2	
Full Title:	Determining the subspecies composition of bean goose harvests in Finland using genetic methods	
Article Type:	Original Article	
Keywords:	Anser fabalis; management; mtDNA; microsatellites; hybridization; sex-biased dispersal	
Corresponding Author:	Johanna Honka, M.Sc University of Oulu Oulu, FINLAND	
Corresponding Author Secondary Information:		
Corresponding Author's Institution:	University of Oulu	
Corresponding Author's Secondary Institution:		
First Author:	Johanna Honka	
First Author Secondary Information:		
Order of Authors:	Johanna Honka	
	Laura Kvist	
	Marja E. Heikkinen	
	Pekka Helle	
	Jeremy B. Searle	
	Jouni Aspi	
Order of Authors Secondary Information:		
Funding Information:	Finnish Game and Fisheries Research Institute (now the Natural Resources Institute Finland)	Not applicable
Abstract:	<p>Management of harvested species is of great importance in order to maintain a sustainable population. Genetics is, however, largely neglected in management plans. Here, we analysed the genetics of the bean goose (<i>Anser fabalis</i>) in order to aid conservation actions for the commonly hunted but declining subspecies, the taiga bean goose (<i>A. f. fabalis</i>). We used mitochondrial DNA (mtDNA) and microsatellites to determine the subspecies composition of the Finnish bean goose harvest, as the hunting bag is thought to comprise two subspecies, the taiga bean goose and the tundra bean goose (<i>A. f. rossicus</i>). The latter subspecies has a more stable or even increasing population size. Other eastern subspecies (<i>A. f. serrirostris</i>, <i>A. f. middendorffii</i>) could additionally be part of the Finnish hunting bag. We estimated genetic diversity, genetic structure and sex-biased gene flow of the different subspecies. Most of the harvested bean geese belonged to the taiga bean goose, whereas most of the tundra bean goose harvest was found to be geographically restricted to south-eastern Finland. The mtDNA data supported strong genetic structure, while microsatellites showed much weaker structuring. This is probably due to the extreme female philopatry of the species. The taiga bean goose had lowered genetic diversity compared to other subspecies, warranting management actions. We also detected <i>A. f. serrirostris</i> mtDNA haplotypes and evidence of interspecific hybridization with two other <i>Anser</i> species.</p>	

1 Johanna Honka, Laura Kvist, Marja E. Heikkinen, Pekka Helle, Jeremy B. Searle and Jouni

2 Aspi

3

4

5 **Determining the subspecies composition of bean goose harvests in Finland**
6 **using genetic methods**

7
8 J. Honka, Genetics and Physiology Unit, University of Oulu, Oulu 90014, Finland

9 Corresponding author

10 email. johanna.honka@oulu.fi

11 L. Kvist, Ecology Unit, University of Oulu, Oulu 90014, Finland

12 M.E. Heikkinen, Genetics and Physiology Unit, University of Oulu, Oulu 90014, Finland

13 P. Helle, Natural Resources Institute Finland, University of Oulu, Oulu 90014, Finland

14 J.B. Searle, Department of Ecology and Evolutionary Biology, Cornell University, Ithaca, NY

15 14853, USA

16 J. Aspi, Genetics and Physiology Unit, University of Oulu, Oulu 90014, Finland

17

18

19 **Acknowledgments**

20 We thank the Finnish Game and Fisheries Research Institute (now the Natural Resources

21 Institute Finland) for providing samples and funding. We thank also Konstantin Litvin for

22 samples, Petri Lampila and Tomas Aarvak for help with morphological examinations and

23 Soile Alatalo for help with laboratory work. We would also like to thank the two anonymous

24 reviewers for their helpful comments on the earlier versions of the manuscript.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

25 **Abstract**

1
2 26 Management of harvested species is of great importance in order to maintain a sustainable
3
4 27 population. Genetics is, however, largely neglected in management plans. Here, we analysed
5
6
7 28 the genetics of the bean goose (*Anser fabalis*) in order to aid conservation actions for the
8
9
10 29 commonly hunted but declining subspecies, the taiga bean goose (*A. f. fabalis*). We used
11
12 30 mitochondrial DNA (mtDNA) and microsatellites to determine the subspecies composition of
13
14 31 the Finnish bean goose harvest, as the hunting bag is thought to comprise two subspecies, the
15
16
17 32 taiga bean goose and the tundra bean goose (*A. f. rossicus*). The latter subspecies has a more
18
19 33 stable or even increasing population size. Other eastern subspecies (*A. f. serrirostris*, *A. f.*
20
21 34 *middendorffii*) could additionally be part of the Finnish hunting bag. We estimated genetic
22
23
24 35 diversity, genetic structure and sex-biased gene flow of the different subspecies. Most of the
25
26
27 36 harvested bean geese belonged to the taiga bean goose, whereas most of the tundra bean
28
29 37 goose harvest was found to be geographically restricted to south-eastern Finland. The mtDNA
30
31
32 38 data supported strong genetic structure, while microsatellites showed much weaker
33
34 39 structuring. This is probably due to the extreme female philopatry of the species. The taiga
35
36 40 bean goose had lowered genetic diversity compared to other subspecies, warranting
37
38
39 41 management actions. We also detected *A. f. serrirostris* mtDNA haplotypes and evidence of
40
41 42 interspecific hybridization with two other *Anser* species.

43
44
45
46 44
47
48 45 **Keywords:** *Anser fabalis*, management, mtDNA, microsatellites, hybridization, sex-biased
49
50
51 46 dispersal

52
53 47
54
55
56
57
58
59
60
61
62
63
64
65

48 **Introduction**

1
2 49 Management of harvested species is necessary to ensure that populations are maintained at a
3
4 50 sustainable level. Sustainable hunting has been defined as ‘the use of wild game species and their
5
6
7 51 habitats in a way and at a rate that does not lead to the long-term decline of biodiversity or hinder
8
9
10 52 its restoration’ (Council of Europe 2007). Among the three conventionally recognised levels of
11
12 53 biodiversity (ecosystem, species and genetic), genetics has largely been neglected in practical
13
14 54 management, as well as in national and international policies (Laikre 2010), especially in relation
15
16
17 55 to hunting. For example, the IUCN (International Union for Conservation of Nature) Red List at
18
19 56 present lacks any genetic criteria (Rivers et al. 2014). Shortage of genetics in practical
20
21
22 57 management is not due to the lack of research or scientific guidelines, but due to failure to
23
24 58 consider genetic issues in management (Frankham 2010). Genetic factors, such as inbreeding
25
26 59 and loss of genetic diversity compromise the viability of populations and may even lead to
27
28
29 60 extinction (Frankham 2005). Harvesting itself may cause genetic changes, such as alteration of
30
31
32 61 population subdivision and loss of genetic variation and local adaptations (Allendorf et al. 2008).
33
34 62 Hence, genetic issues should be incorporated into management of harvested species in order to
35
36 63 manage populations properly (Waples and Gaggiotti 2006; Palsbøll et al. 2007).

37
38
39 64 Only a few studies have focused on the incorporation of genetic aspects into management
40
41 65 plans. Moyle et al. (2003) studied the species recovery plans (1977-1998) in the United States
42
43
44 66 and they concluded that genetics had only a minor role and that the understanding of how
45
46 67 genetics could be used to aid the species recovery was limited. However, in some cases genetics
47
48
49 68 has been successfully incorporated into management actions. One famous example is the genetic
50
51 69 restoration of the endangered Florida panther (*Puma concolor coryi*; U.S. Fish and Wildlife
52
53 70 Service 2008; Johnson et al. 2010). The Florida panthers had low genetic variation, which was
54
55
56 71 an indication of inbreeding that may have led to several defects, such as poor sperm quality,
57
58 72 cryptorchidism (testicles not descending), kinked tail and cowlick on the back. Genetic rescue
59
60
61
62
63
64
65

73 with the translocation of eight Texas female pumas (*P. c. stanleyana*) led to increased population
74 numbers and reduced the incidence of inbreeding defects (Johnson et al. 2010).

75 A process readily revealed by genetic analysis is hybridization, as defined as the
76 interbreeding of individuals from genetically distinct populations (Short 1969). Hybridization is
77 a serious conservation problem as demonstrated in many populations and species of plants and
78 animals (Todesco et al. 2016). The risk of extinction by hybridization is increased by human
79 activities such as translocations, husbandry and habitat disturbance, especially in the absence of
80 reproductive barriers and when there is introgression (gene flow from one population to the other
81 as a result from hybridization) to the rare species (Todesco et al. 2016). As an example,
82 hybridization of the endangered red wolf (*Canis rufus*) with coyotes (*C. latrans*) has been
83 considered as one of the greatest threats to the red wolf (Gese et al. 2015). The red wolf numbers
84 declined due to land use changes, which allowed the coyotes to invade their range and hybridize
85 with the red wolves, leading to a loss of almost all of the red wolf populations due to genetic
86 mixing. The red wolf adaptive management plan (U.S. Fish and Wildlife Service 2013) has been
87 successful at reducing the nuclear introgression of coyote genes into the red wolf (Gese et al.
88 2015).

89 Many European migrating waterfowl populations are of management concern and the EU
90 Birds Directive and the African-Eurasian Waterbird Agreement (AEWA) provide the legal
91 framework for sustainable management of migratory waterfowl populations. However,
92 International Single Species Action Plans (ISSAPs) by AEWA fail to incorporate genetics into
93 management actions and goals for the conservation of geese (Robinson and Colhoun 2006;
94 Cranswick et al. 2012; Madsen and Williams 2012; Stroud et al. 2012; Marjakangas et al. 2015),
95 except in the case of the lesser white-fronted goose (*Anser erythropus*; Jones et al. 2008). The
96 ISSAP of the lesser white-fronted goose advocates development of genetic assessments and a
97 strategy for genetic management and minimisation of interspecific introgression due to captive

98 breeding programs (Jones et al. 2008). Among goose species that are currently exploited by man,
99 genetic methods have only been employed for the Canada goose (*Branta canadensis*), with the
100 composition of the hunting bag genetically assessed (Inman et al. 2003; Scribner et al. 2003;
101 Shorey et al. 2007; Mylecraine et al. 2008) and included in the Canada goose management plans
102 (Canada Goose Committee 2008; Pacific Flyway Council 2015).

103 The bean goose (*Anser fabalis*) is currently divided into four subspecies: *A. f. fabalis*, *A. f.*
104 *rossicus*, *A. f. serrirostris* and *A. f. middendorffii* (Fig. 1). However, taxonomy of the bean goose
105 has been controversial for a long time with numerous changes (Delacour 1951; Sangster and
106 Oreel 1996; Ruokonen and Aarvak 2011). Traditionally, five subspecies were recognised
107 including also *A. f. johanseni* (Delacour 1951), whose validity as a subspecies has later been
108 rejected by several authors (Burgers et al. 1991; Sangster and Oreel 1996; Ruokonen and Aarvak
109 2011). Historically, the pink-footed goose (*A. brachyrhynchus*) was also considered as a
110 subspecies of the bean goose (Delacour 1951). More recently, Sangster and Oreel (1996)
111 suggested that there are two species *A. fabalis* (including *A. f. fabalis* and *A. f. middendorffii*) and
112 *A. serrirostris* (including *A. f. rossicus* and *A. f. serrirostris*). Two species was also proposed by
113 Ruokonen et al. (2008) based on the mtDNA sequences, but with differing composition. *A.*
114 *fabalis* was suggested to include three subspecies *A. f. fabalis*, *A. f. rossicus* and *A. f. serrirostris*
115 whereas *A. middendorffii* forms another species, the Middendorf's goose, as it was clearly
116 differentiated from the bean geese (Ruokonen et al. 2008). Further, Ruokonen et al. (2008)
117 classified *A. brachyrhynchus* as a separate species based on the mtDNA, but a recent exon-based
118 phylogenomics study identified a sister-species relationship of *A. brachyrhynchus* and *A. f.*
119 *rossicus* (Ottenburghs et al. 2016a). This incongruence between different genetic markers could
120 be due to very recent speciation that still can be seen as incomplete lineage sorting and/or
121 hybridization in the bean goose-pink-footed goose complex (Ruokonen et al. 2000; Ottenburghs

122 et al. 2016a). However, the study of Ottenburghs et al. (2016a) did not include all the bean goose
123 subspecies, leaving the evolutionary relationships of the complex still unresolved.

124 The bean goose subspecies are grouped into breeding forms (Delacour 1951) that inhabit
125 different habitats in Fennoscandia and Russia (Fig. 1). The taiga breeding forms (*A. f. fabalis*, *A.*
126 *f. middendorffii*) inhabit open or wooded mires, small lakes, ponds and streams (Nilsson et al.
127 1999) whereas the tundra breeding forms (*A. f. rossicus*, *A. f. serrirostris*) inhabit open tundra,
128 usually near lakes or rivers (Van den Bergh 1999). The breeding forms differ slightly in body
129 size and shape, bill morphology and coloration and plumage colour but due to large individual
130 variation in morphology, the visual identification of each subspecies is challenging (Delacour
131 1951). Consequently, the breeding forms or the subspecies are not identified in goose counts or
132 in hunting statistics.

133 Most of the European goose populations are currently expanding, while only the lesser-
134 white fronted goose is showing a long-term decline and the red-breasted goose (*B. ruficollis*) and
135 the taiga bean goose (*A. f. fabalis*) are showing short-term declines (Fox et al. 2010). The most
136 recent population estimate for *A. f. fabalis* is 40,000-50,000 individuals at the end of the non-
137 breeding season (Wetlands International 2016) which is less than half of the estimated 90,000-
138 100,000 individuals in the 1990s (Nilsson et al. 1999). The taiga bean goose population was still
139 expanding in 1970-1990 (Madsen 1991), but has been decreasing since then (Fox et al. 2010).
140 The population trend of the western tundra bean goose *A. f. rossicus* has been stable (Fox et al.
141 2010) and estimated to be approximately 550,000 individuals (Wetlands International 2016). The
142 decline of the taiga bean goose is of great management concern as the species is hunted
143 throughout its range (Fig. 1) except in Great Britain, Norway, the Netherlands and Belgium, with
144 thousands of taiga bean geese hunted in Finland, Sweden and Russia, and hundreds in Denmark
145 (Hirschfeld and Heyd 2005).

146 Here, we aim to 1) determine the ratio between the declining *A. f. fabalis* and the stable *A.*
1
2 147 *f. rossicus* in the Finnish hunting bag, 2) examine the possible presence of the eastern breeding
3
4 148 taiga- and tundra bean geese in the hunting bag and search for possible hybridization between
5
6
7 149 geese species and 3) provide estimates of genetic diversity, genetic structure and sex-biased gene
8
9
10 150 flow for bean goose subspecies. We use mitochondrial DNA control region sequences and
11
12 151 microsatellites to determine subspecies composition of the bean goose hunting bag in Finland.
13
14 152 The mtDNA control region has been shown to separate well the different subspecies (Ruokonen
15
16
17 153 et al. 2008) and we will also evaluate the usefulness of microsatellites in the subspecies
18
19 154 identification.

20
21
22 155

23 24 156 **Material and methods**

25 26 157 **Sampling and DNA extraction**

27
28
29 158 The Finnish bean goose specimens ($n=103$) consisted of wings collected by hunters during the
30
31 159 legal hunting seasons in years 2010-2013 (2010 $n=9$; 2011 $n=26$; 2012 $n=64$ and 2013 $n=4$). In
32
33
34 160 2010 the hunting season began on 10 September, in 2011 on 17 September, in 2012 on 26
35
36 161 September and in 2013 on 10 October, in southern and central parts of Finland. In northern and
37
38
39 162 eastern parts (Lapland and Kainuu) the season began earlier. Each year the hunting season
40
41 163 continued until the end of December. Since 2014, hunting of the bean goose has been completely
42
43 164 forbidden in Finland. We also included Norwegian ($n=8$), Russian ($n=39$) and Finnish ($n=8$)
44
45
46 165 samples of known breeding origin from years 1997-2006 (Fig. 1). Most of the latter samples
47
48
49 166 ($n=41$) were included in a study by Ruokonen et al. (2008). In addition, we used mitochondrial
50
51 167 control region sequences from GenBank published by Ruokonen et al. (2000, 2008): EU186805–
52
53 168 EU186812 and AF159951 (*A. f. fabalis* haplotypes FAB1a, FAB1b and FAB3, *A. f. rossicus*
54
55
56 169 haplotypes ROS2a and ROS2b, *A. f. serrirostris* haplotypes SER1a and SER1b and *A.*
57
58 170 *middendorffii* haplotypes MID1 and MID5) as well as partial control region (219 bp) sequences
59
60
61
62
63
64
65

171 (Ruokonen et al. 2008) EU186813–EU186828 (*A. f. fabalis* FAB1, FAB3 and FAB6, *A. f.*
172 *rossicus* ROS2, ROS3 and ROS4, *A. f. serrirostris* SER1, SER2 and SER3 and *A. middendorffii*
173 MID1, MID2, MID4, MID5 and MID6). We extracted DNA using a DNeasy Blood and Tissue
174 Sample Kit (Qiagen) according to manufacturer’s instructions from 5 µl of blood (4 samples) or
175 about 4 mg of muscle (all the rest). The samples with known breeding origin had their DNA
176 extracted as in Ruokonen et al. (2008).

177
178 Mitochondrial DNA sequencing
179 We amplified the whole tRNA_{glu} gene and almost the whole mitochondrial control region (11
180 bp from 3’ end was excluded as in Ruokonen et al. 2008) that has been shown to distinguish the
181 four bean goose subspecies (Ruokonen et al. 2008). The 1235 bp sequence was amplified in two
182 fragments with primer pairs L16642/H411-AL and L334-AL/H1248 (Ruokonen et al. 2000), that
183 were designed to contain mismatches to Numts (nuclear sequences of mitochondrial origin;
184 Lopez et al. 1994). We performed PCR in 20 µl reaction volumes using Phusion High-Fidelity
185 DNA Polymerase (Thermo Fisher Scientific) and 50-100 ng of template-DNA. For PCRs with
186 primers L16642/H411-AL, the thermal profile consisted of 98 °C for 30s, followed by 30 cycles
187 of 98 °C for 10s, 52 °C for 20s and 72 °C 15s with a final extension of 72 °C for 10min. For
188 PCRs with primers L334-AL/H1248, we used the same thermal profile except for primer
189 annealing temperature of 59 °C for 30s and synthesis for 30s. Double-stranded sequencing of the
190 PCR products with the PCR primers was performed using BigDye Terminator v.3.1 (Applied
191 Biosystems) and the reactions were run on an ABI 3730 (Applied Biosystems). We aligned and
manually edited sequences using BioEdit 7.2.5 (Hall 1999).

193
8

194 Microsatellite genotyping

1
2 195 For the microsatellite analysis, we chose 20 polymorphic loci originally designed for a closely
3
4
5 196 related species, the pink-footed goose: Abra2, Abra3, Abra4, Abra5, Abra7, Abra9, Abra10,
6
7 197 Abra12, Abra14, Abra15, Abra19, Abra23, Abra24, Abra29, Abra30, Abra35, Abra39, Abra43,
8
9
10 198 Abra49 and Abra68 (Table 1; Noreikiene et al. 2012). The forward primers were fluorescently
11
12 199 labelled with VIC, PET, FAM or NED. The microsatellite amplification was performed in two
13
14 200 multiplexes using Type-it Microsatellite PCR Kit (Qiagen) in 10 µl volumes according to
15
16 201 manufacturer's instructions. The annealing temperature was set to 60 °C (see Noreikiene et al.
17
18
19 202 2012). We performed the fragment analysis with an ABI 3730 and scored alleles with
20
21
22 203 GeneMapper 5 (Applied Biosystems). We amplified all the samples twice to assess genotyping
23
24 204 error between the two runs by calculating the number of mismatched genotypes divided by the
25
26 205 number of reactions (Hoffmann and Amos 2005).

27
28
29 206
30
31 207 Mitochondrial DNA analysis

32
33
34 208 We estimated the genetic variation by calculating the number of haplotypes, haplotype diversity
35
36 209 and nucleotide diversity with DnaSP v.5 (Librado and Rozas 2009). To estimate the
37
38
39 210 differentiation between the subspecies, we calculated ϕ_{ST} values using the Tamura-Nei genetic
40
41 211 distance (Tamura and Nei 1993) and alpha value 0.05 (significance tested with 10 000
42
43
44 212 permutations) with Arlequin 3.5.1.3 (Excoffier and Lischer 2010). We conducted AMOVA
45
46 213 (analysis of molecular variance; Excoffier et al. 1992), as implemented in Arlequin 3.5.1.3, to
47
48
49 214 partition the mtDNA diversity among subspecies derived from the phylogenetic analyses.
50
51 215 Demographic and spatial population expansion of each subspecies was examined by calculating
52
53 216 Tajima's D (Tajima 1989), Fu's F_s (Fu 1997) and R_2 (Ramos-Onsins and Rozas 2002) with
54
55
56 217 coalescent simulations using the DnaSP v.5 and the mismatch distribution using Arlequin
57
58 218 3.5.1.3.

219 We used MEGA 6.06 (Tamura et al. 2013) to choose the appropriate DNA substitution
220 model and selected the HKY+G model (Hasegawa et al. 1985) as both AIC (Akaike Information
221 Criteria; 4486) and BIC (Bayesian Information Criterion; 5029) values supported this model
222 with an alpha value of 0.05. We constructed a phylogenetic tree of the haplotypes inferred with
223 DnaSP v.5 using MrBayes v.3.2.2. (Ronquist et al. 2012) with four incrementally heated MCMC
224 chains for 1,000,000 generations, 100 as the sampling frequency, 0.05 as the Temp parameter
225 and discarded 25% of the first trees as a burn-in, using otherwise the default parameters. The
226 average standard deviation of split frequencies was 0.004. We also evaluated the convergence of
227 the runs using Tracer v1.6 (Rambaut et al. 2014) by checking that the effective sample sizes
228 (ESS) were >200 for all estimated parameters (> 3000 in our runs). The consensus tree was
229 visualised using FigTree v1.4.2 (Rambaut 2006-2014). We also constructed a Maximum
230 Likelihood tree with MEGA 6.06 using the HKY+G model with 1000 bootstrap replicates and
231 using default parameters otherwise (the tree we present excluded gaps in the analysis; similar
232 results were obtained when gaps were included). We constructed the trees using the 1235 bp
233 control region sequences and using only the 219 bp hypervariable part of the control region in
234 order to identify also haplotypes previously defined only by the shorter fragment (see Ruokonen
235 et al. 2008). We constructed a Median-Joining network (Bandelt et al. 1999) using the program
236 PopART (Leight and Bryant 2015) with ϵ set to zero. GenBank sequences of the mtDNA control
237 region from the greylag goose (*A. anser*; GenBank Accession number AF159961), the greater
238 white-fronted goose (*A. albifrons*; AF159958) and the pink-footed goose (AF159952 and
239 AF159953) were used as outgroups.

241 Microsatellite analysis

242 We used Micro-Checker (Van Oosterhout et al. 2004) to search for null alleles and genotyping
243 errors in the data and estimated the frequency of null alleles with the program FreeNA (Chapuis

244 and Estoup 2007). The samples were classified to subspecies based on their mitochondrial
245 sequences and the following analyses were performed to these subspecies-groups. Deviations
246 from Hardy-Weinberg Equilibrium (HWE) for each loci (Fisher exact test), Linkage
247 Disequilibrium (LD) for each pair of loci in each population, R_{ST} (Slatkin 1995) values, observed
248 (H_O) and expected (H_E) heterozygosities were estimated with Genepop 4.3 (Rousset 2008) and
249 number of alleles, allele richness and pairwise F_{ST} (Wright 1951) values were estimated with
250 FSTAT 2.9.3.2 (Goudet 1995). The effect of mutations on population differentiation was
251 assessed with a permutation test (1000 permutations) implemented in the program SPAGeDi 1.5
252 (Hardy and Vekemans 2002) by testing if $F_{ST}=R_{ST}$ (Hardy et al. 2003). Sequential Bonferroni
253 correction was applied to Hardy-Weinberg tests, F -statistics and linkage equilibrium (Rice
254 1989). The unbiased probability of identity (P_{ID}) and the probability of identity of siblings (P_{ID}
255 Sib) were estimated with the program Gimlet v.1.3.3 (Valière 2002).

256 In addition, population structure was studied using the program Genetix 4.05 (Belkhir et al.
257 2004) for a Factorial Correspondence Analysis (FCA) and the program Structure v.2.3.4
258 (Pritchard et al. 2000; Falush et al. 2003) for a clustering analysis. Structure was run first without
259 prior information of populations, with a run length of 500,000 and burn-in 50,000 with the
260 number of possible clusters (K) set from 1–7 and 8 iterations for each. The ancestry model was
261 set to admixture and correlated allele frequencies were used. We inferred the most likely number
262 of clusters on the basis of ΔK values estimated using the *ad hoc* approach of Evanno et al. (2005)
263 implemented in Structure Harvester (Earl and vonHoldt 2012). We also ran Structure using prior
264 population information (Usepopinfo) based on subspecies assignment from the mtDNA results.
265 The run length was set to 1,000,000, burn-in to 100,000 and K=4 according to subspecies
266 number (*A. f. fabalis*, *A. f. rossicus*, *A. f. serrirostris* and *A. f. middendorffii*) using population
267 information in the ancestry model (Migprior=0.05) together with correlated allele frequencies.

269 Effects of year and wind direction

270 We tested for the difference in the numbers of *A. f. fabalis* and *A. f. rossicus* in the hunting bag
271 between the years 2010–2012 with a χ^2 -test, comparing the impact of easterly and northerly
272 winds each year. We obtained wind direction data from the Finnish Meteorological Institute
273 (2015) as determined in September and October each year for south-eastern Finland (Virolahti
274 and Lappeenranta), where most of the *A. f. rossicus* were harvested. Year 2013 was excluded
275 from the analyses due to a low sample size of geese ($n=4$).

277 Sex-biased dispersal

278 We estimated sex-biased gene flow for different subspecies by calculating the differentiation for
279 males and females and the ratio of male to female gene flow using a F_{ST} -based method suggested
280 by Hedrick et al. (2013; see also Hedrick et al. 2015) using their equations 7a and 7b. We also
281 estimated the sex-biased dispersal between Canada goose populations using the same method
282 with the data of F_{ST} values in Mylecraine et al. (2008) in order to conduct an interspecific
283 comparison.

285 **Results**

286 Mitochondrial DNA

287 We obtained the tRNA_{glu} and almost the whole control region sequence (1235 bp) which
288 includes the hypervariable region (219 bp) from 121 bean geese of which 96 (out of 103) were
289 the Finnish hunted geese of unknown breeding origin and 25 (out of 55) had a known breeding
290 origin. We obtained only partial or no sequence from the rest of the samples and did not include
291 these in further analyses. There were altogether 27 haplotypes of which 18 were from the hunted
292 geese. We did not find any sequence length variation in the bean goose, however in

293 Anseriformes a C-stretch in the 5'-end can form a hairpin structure making the interpretation of
1
2 294 the number of cytosines challenging.
3

4
5 295 The haplotype diversity was the highest in *A. f. middendorffii* and the second highest in *A.*
6
7 296 *f. serrirostris* (Table 1) when the subspecies were compared. The lowest haplotype diversity was
8
9 297 found in *A. f. fabalis*. Nucleotide diversity was also the highest in *A. f. middendorffii* and the
10
11 298 lowest in *A. f. fabalis* (Table 1). 83% of the observed variation was explained by among
12
13 299 subspecies and 17% by within subspecies variation in the AMOVA results. In all the subspecies
14
15 300 comparisons the pairwise Φ_{ST} values were high (0.68-0.86; all $P < 0.001$; Table 2). *A. f.*
16
17 301 *middendorffii* was the most differentiated from the rest of the subspecies and *A. f. rossicus* and *A.*
18
19 302 *f. serrirostris* were the least differentiated from each other. Tajima's D and Fu's F_s were
20
21 303 negative in all subspecies except in *A. f. serrirostris*, but significant only in *A. f. rossicus* (Table
22
23 304 1). Mismatch Distribution (MD) analysis and R_2 statistics with coalescent simulation indicated
24
25 305 no population expansion for any of the populations (sum of squared deviation and R_2 , all $P >$
26
27 306 0.05), but the raggedness-value for *A. f. rossicus* was significant (0.03, $P < 0.05$; Table 1).
28
29
30
31
32
33

34 307 Phylogenetic trees constructed by Bayesian and Maximum Likelihood methods produced
35
36 308 similar tree topologies, though the Bayesian posterior probabilities for the different branches
37
38 309 were higher than the Maximum Likelihood bootstrap support (Online resource 1). The *A.*
39
40 310 *brachyrhynchus*-, *A. f. middendorffii*- and *A. f. fabalis/A. f. rossicus/A. f. serrirostris*-groups were
41
42 311 clearly separated in the tree (posterior probabilities 1.00, 0.99 and 0.95, respectively), with the
43
44 312 latter group further separated into *A. f. fabalis* and *A. f. rossicus/A. f. serrirostris*-groups. *A. f.*
45
46 313 *middendorffii* was clearly separated as its own cluster, with high posterior probabilities (0.99)
47
48 314 and bootstrap values (83%), as well as *A. f. fabalis* (1.00/82%), but the support for the *A. f.*
49
50 315 *rossicus* and *A. f. serrirostris* groups was much lower (0.69/38%) (Online resource 1). The
51
52 316 topology of the median-joining network was in accordance with the phylogenetic results with the
53
54 317 haplotype Fa3 being the most common in *A. f. fabalis* and ROS2a in *A. f. rossicus* (Fig. 2).
55
56
57
58
59
60
61
62
63
64
65

318 One individual from the Finnish bean geese hunting bag carried the mtDNA sequence of
1
2 319 the pink-footed goose (haplotype Br1, Fig. 2) and another individual carried the mtDNA of the
3
4 320 white-fronted goose (haplotype A11, Fig. 2). Three bean geese from Valdak Norway, sampled in
5
6
7 321 2003 represented the subspecies *A. f. rossicus* according to their mtDNA. In the Finnish bag,
8
9 322 there was three of the four bean goose subspecies present on the basis of their mtDNA: *A. f.*
10
11 323 *fabalis* 52% ($n=53$), *A. f. rossicus* 44% ($n=45$) and *A. f. serrirostris* 2% ($n=2$) (Online resource
13
14 324 2). The remaining 2% were the two individuals with the mtDNA of other geese species. The
15
16 325 hunting locations of the subspecies varied geographically, with *A. f. fabalis* hunted throughout
17
18 326 Finland but *A. f. rossicus* mostly in south-eastern Finland with only few individuals hunted
19
20 327 outside that region (Fig. 3).
21
22
23
24 328

26 329 Microsatellites

29 330 Two loci (Abra3 and Abra4) failed to amplify, one locus (Abra49) showed ambiguous results
30
31 331 and one locus (Abra35) contained null alleles in all subspecies studied and therefore these four
32
33 332 loci were excluded, leaving 16 loci for further analyses. We succeeded in genotyping a total of
34
35 333 153 samples of which 103 (out of 103) were from the Finnish hunted geese and 50 (out of 55)
36
37 334 were from the geese with a known breeding origin. Genotyping error between the two
38
39 335 amplifications of the same samples averaged over loci was low for both the Finnish hunted geese
40
41 336 (0.007) and for the geese of known breeding origin (0.026). The program Micro-Checker
42
43 337 suggested null alleles (at a frequency of 0.00–0.24) at a few loci which deviated from the Hardy-
44
45 338 Weinberg equilibrium ($P < 0.05$). These loci were however included in the further analysis
46
47 339 because the deviation from Hardy-Weinberg was probably due to population structure;
48
49 340 indications for null alleles were not constant across the loci or subspecies (except for the
50
51 341 excluded Abra35) and the frequency of null alleles was low in most loci. Estimates of F_{ST} using
52
53 342 the data corrected with FreeNA did not differ significantly from the uncorrected values, so the
54
55
56
57
58
59
60
61
62
63
64
65

343 existence of any null alleles did not bias our results. We did not find any Linkage Disequilibrium
1
2 344 after Bonferroni correction. Numbers of alleles were highest in *A. f. fabalis* and *A. f. rossicus*, but
3
4 345 these subspecies had the largest sample sizes (Table 1). The allelic richness, which takes into
5
6
7 346 account the differences in the sample sizes did not vary much between the subspecies (3.3-3.8),
8
9
10 347 and neither did the estimates for heterozygosity ($H_O= 0.43-0.50$; $H_E= 0.51-0.57$; Table 1). P_{ID}
11
12 348 values varied from 0.07 to 0.91 and $P_{ID\ Sib}$ values varied from 0.38 to 0.96 (Table 1).

14 349 The microsatellite markers did not show differentiation between the subspecies. Also the
15
16
17 350 individuals with the mtDNA derived from another species clearly fell inside the intraspecific
18
19 351 variation of the bean goose. The pairwise F_{ST} values were very low (≤ 0.03), especially when
20
21
22 352 compared to the ϕ_{ST} values from the mtDNA (Table 2). The R_{ST} values were higher than the F_{ST}
23
24 353 values except for the *A. f. fabalis* – *A. f. rossicus* pair (Table 2). A permutation test indicated that
25
26
27 354 R_{ST} values were significantly higher ($P < 0.05$) than F_{ST} values only in two loci (Abra9 and
28
29 355 Abra12), which indicates that only these two loci evolve under a strict stepwise mutations model.
30
31
32 356 Thus, we used only F-statistics with the microsatellite loci. Factorial Correspondence Analysis
33
34 357 (FCA) also showed high genetic similarity between the subspecies, although some slight
35
36 358 differentiation was seen between *A. f. fabalis* and *A. f. rossicus* (Fig. 4). Structure analysis
37
38
39 359 without prior population information gave $K=3$ as the most probable number of genetic clusters,
40
41 360 but all individuals showed a high amount of admixture and no clear clustering (Online resource
42
43
44 361 3a). However, when we used the subspecies assignment based on the mtDNA results as prior
45
46 362 population information ($K=4$), Structure clustered the individuals clearly according to their
47
48
49 363 subspecies (Online resource 3b). All the individuals belonged to their subspecies with high
50
51 364 likelihood (80-90%) except one *A. f. rossicus* individual that was admixed with *A. f. fabalis* (58%
52
53 365 *rossicus* and 42% *fabalis*). The two geese with the mtDNA from a different species clearly
54
55
56 366 belonged to bean goose according to nuclear genotypes (Online resource 3b).

58 367

368 Effects of year and wind direction

369 The frequency of the harvested subspecies varied between years (Table 3) with about 30% more
370 *A. f. fabalis* in 2011 than in other years ($\chi^2 = 9.07$, $P < 0.05$). In 2010 and 2012 there were slightly
371 more *A. f. rossicus* than *A. f. fabalis*. The prevailing wind direction varied between the years
372 ($\chi^2 = 267.8$, $P < 0.01$) with year 2012 having more easterly winds than compared to other years
373 (Online resource 4). The frequencies of *A. f. rossicus* as well as the easterly winds were higher in
374 2012 than in 2011. This indicates that the easterly winds shifted the migration of more *A. f.*
375 *rossicus* individuals into Finland in autumn while prevailing northerly winds increased the
376 proportion of *A. f. fabalis* in the hunting bag.

377
378 Sex-biased dispersal
379 Gene flow between the subspecies was much higher in the bean goose males than in the females
380 and the ratio of gene flow between males and females (m_m/m_f) varied between (57.2-316; Table
381 4), with an average of 122. In the Canada goose, the m_m/m_f ratio varied from -0.35 to 17.40
382 between populations, with an average of 4.47. Compared to the bean goose, the Canada goose
383 showed much lower m_m/m_f ratios, suggesting less sex-biased dispersal.

384
385 **Discussion**

386 Composition of the hunting bag
387 The Finnish bean goose harvests consisted mainly of the subspecies *A. f. fabalis* and *A. f.*
388 *rossicus* as expected, since *A. f. fabalis* is the main subspecies breeding in Finland and *A. f.*
389 *rossicus* is a regular passage migrant in Finland. Based on our results and the previous study by
390 Ruokonen et al. (2008), we confirmed that the mtDNA-based classification of individuals to
391 subspecies is powerful. The microsatellite data gave less clear results than mtDNA and the
392 subspecies assignment could not be performed based on microsatellites alone. However, when

393 subspecies information obtained from mtDNA was used, the microsatellite data fit well with the
1
2 394 subspecies assignments, supporting that the subspecies form coherent taxonomic entities.
3

4
5 395 The proportion of different subspecies fluctuated between years with *A. f. rossicus*
6
7 396 harvested more than *A. f. fabalis* in most years but, on the whole, more *A. f. fabalis* were
8
9 397 harvested. The fluctuation could be partly explained by the prevailing wind directions. When the
10
11 398 easterly winds dominate, they shift the migration route of *A. f. rossicus* to the south-eastern
12
13 399 Finland from Russia, whereas when northerly winds dominate, the migration route of *A. f.*
14
15 400 *rossicus* stays mostly in Russia (Toivainen et al. 2014). This would result in more *A. f. rossicus*
16
17 401 being hunted in Finland when easterly winds prevail during the migration time. It has been
18
19 402 observed that winds shift the migration routes in other goose species as well (for example in the
20
21 403 Brent goose *B. bernicla* and in the barnacle goose *B. leucopsis*; Green 2001). However, our
22
23 404 results need further confirmation as sample sizes were rather low.
24
25
26
27
28

29 405 The subspecies composition in the hunting bag varied geographically, with *A. f. rossicus*
30
31 406 hunted almost solely from south-eastern Finland along the Russian border, while *A. f. fabalis* was
32
33 407 hunted evenly over the whole Finland (Fig. 3). This was predicted, as *A. f. fabalis* breeds in
34
35 408 northern and central Finland and passes through the Åland archipelago located between Finland
36
37 409 and Sweden to staging areas in southern Sweden (Nilsson 2011). In addition, Russian *A. f.*
38
39 410 *fabalis* migrates also through Finland (Nilsson 2011). On the contrary, *A. f. rossicus* migrates
40
41 411 along the eastern border of Finland via the Baltic countries to Central and Eastern Europe (Van
42
43 412 den Bergh 1999). Interestingly, one *A. f. rossicus* individual was hunted in the Finnish Lapland
44
45 413 and thus could originate from the quite recently reported *A. f. rossicus* population in the
46
47 414 Norwegian Finnmark (Aarvak and Øien 2009) that migrates along the coast of Sweden (De Jong
48
49 415 et al. 2013). Also a few *A. f. rossicus* individuals were harvested at or near the Finnish west
50
51 416 coast. These birds could have used some alternative migration route or have wandered off the
52
53 417 main migration routes.
54
55
56
57
58
59
60
61
62
63
64
65

418 In addition to these two main subspecies, two eastern tundra bean geese *A. f. serrirostris*
1
2 419 mtDNA haplotypes were found among the hunted individuals. However, according to Ruokonen
3
4 420 et al. (2008), several *A. f. serrirostris* haplotypes are found also in *A. f. rossicus*, thus the two
5
6
7 421 individuals carrying *A. f. serrirostris* haplotypes might actually represent *A. f. rossicus* (see also
8
9
10 422 Fig. 2), or they could as well be hybrids. It is possible that *A. f. serrirostris* is an occasional
11
12 423 wanderer to Finland, providing an opportunity for hybridization, even though there are no
13
14 424 previous reports of *A. f. serrirostris* in Finland. However, as the resolution in the microsatellites
15
16
17 425 was not sufficient to separate these two subspecies, the identity of these birds remains unclear.

18
19 426 We found one bird with mtDNA of the pink-footed goose and another with mtDNA of the
20
21
22 427 greater white-fronted goose. The microsatellites indicated that these individuals are bean geese
23
24 428 (Online resource 3b). Hence, this implies inter-specific hybridization and introgression of
25
26
27 429 mtDNA to bean goose from other goose species. The bird with the pink-footed goose mtDNA
28
29 430 looked morphologically like *A. f. rossicus*, except for yellower feet and bill than a normal bean
30
31
32 431 goose (Tomas Aarvak, personal communication). The wing of the bird with the greater white-
33
34 432 fronted goose mtDNA looked like a young greater white-fronted goose due to its grey colour and
35
36
37 433 lack of clear white fringes of primary feathers typical of the bean goose (Petri Lampila, personal
38
39 434 communication). These unusual morphological features suggest that these two birds were of
40
41 435 hybrid origin. However, we did not have microsatellite data from pink-footed or white-fronted
42
43
44 436 geese to confirm this result. In addition, the usage of microsatellite loci designed for another,
45
46 437 although closely related, species could have limited the effectiveness of detecting hybrid
47
48
49 438 individuals.

50 51 439 52 53 440 Genetic diversity, genetic structure and gene flow

54
55
56 441 The nuclear diversity in all subspecies, measured as observed heterozygosity, was low (0.43-
57
58 442 0.50) compared for example to the greater white-fronted goose (0.67; Ruokonen et al. 2007).

443 However, it was at the same level as observed in the lesser white-fronted goose (0.51) that has
1
2 444 been strongly declining in population size (Ruokonen et al. 2007). Mitochondrial haplotype
3
4 445 diversities in *A. f. rossicus*, *A. f. serrirostris* and *A. f. middendorffii* were higher ($h = 0.68 - 0.86$)
5
6
7 446 and nucleotide diversities lower ($\pi = 0.001 - 0.002$) compared to several other geese (e.g. lesser
8
9
10 447 white-fronted goose, $h = 0.37 - 0.53$, $\pi = 0.003$, pink-footed goose, $h = 0.51$, $\pi = 0.003$;
11
12 448 Ruokonen et al. 2004, 2005). This could have resulted from a population growth after a past
13
14 449 bottleneck (Grant and Bowen 1998). However, *A. f. rossicus* was the only subspecies showing
15
16
17 450 signs of past population growth also by Tajima's D , Fu's F_s and the raggedness index and is the
18
19 451 only population not in decline at present. *A. f. fabalis*, on the other hand, had lower haplotype
20
21
22 452 and nucleotide diversities ($h = 0.582$, $\pi = 0.00103$) than the other subspecies, suggesting a
23
24 453 possibility of a relatively recent bottleneck (Grant and Bowen 1998). The mitochondrial diversity
25
26
27 454 of *A. f. fabalis* was close to the levels observed in other geese species that are declining or have
28
29 455 had historically low population sizes (such as the above mentioned lesser white-fronted and the
30
31 456 pink-footed goose).

32
33
34 457 The haplotype network clearly supported clustering of the bean goose into three separate
35
36 458 groups: *middendorffii*, *fabalis* and a group including *rossicus* and *serrirostris* (Fig. 2).
37
38
39 459 Divergence between the subspecies measured by the pairwise F_{ST} (0.01-0.03) or R_{ST} (0.01 –
40
41 460 0.07) values of microsatellite data (Table 2) was much lower than from the mtDNA (ϕ_{ST} : 0.68 –
42
43
44 461 0.86). The level of divergence in the bean goose microsatellites is comparable to values obtained
45
46 462 from other goose species, for example the pairwise F_{ST} values between two wild populations of
47
48
49 463 the lesser white-fronted goose was 0.01 (Ruokonen et al. 2007) and between populations of the
50
51 464 Canada goose from 0.002 to 0.05 (Mylecraine et al. 2008).

52
53 465 This discrepancy in the amount of differentiation estimated from the two types of markers
54
55
56 466 can partly be explained by differing effective population sizes of these markers, as mtDNA has
57
58 467 four times smaller effective size than microsatellites. However, the extremely strong philopatry
59
60
61
62
63
64
65

468 in females can also have a great effect (Zink and Barrowclough 2008). When females return to
1
2 469 nest at their natal sites, geographical structure is found in the maternally inherited mtDNA, but
3
4 470 gene flow through males inhibits structuring in nuclear loci (Zink and Barrowclough 2008). We
5
6
7 471 detected up to 300 times greater gene flow in males than in females and this seems to explain
8
9
10 472 most of the difference between markers. This amount of sex-biased gene-flow is much larger
11
12 473 than what we observed by performing the same calculations for the Canada goose (up to 17
13
14 474 times greater). Evidence of the sex-biased dispersal has been found also in the lesser-white
15
16 475 fronted goose (Ruokonen et al. 2010) and the greylag goose (Nilsson and Persson 2001) but not
17
18
19 476 in all goose species (e.g. in the lesser snow goose, *A. caerulescens*; Avise et al. 1992). The strong
20
21
22 477 female philopatry could make the local taiga bean goose populations especially vulnerable to
23
24 478 overharvesting, as local populations are not readily re-colonised after local extinction, due to the
25
26 479 female site fidelity (Marjakangas et al. 2015).

27
28
29 480
30
31 481 Hybridization
32
33
34 482 Ducks and geese (Anseriformes) show the greatest propensity to hybridization in birds, with over
35
36 483 40% of the species doing so (Grant and Grant 1992; Ottenburghs et al. 2016b). For the declining
37
38
39 484 taiga bean goose, hybridization can become a major threat as it brings genes from other species
40
41 485 into the taiga bean goose. In geese, pair bonding takes place during the winter or early in the
42
43
44 486 spring (Rohwer and Anderson 1988). In the winter, the geese are highly gregarious and may
45
46 487 form mixed flocks with other goose species. Sometimes this may lead them to form inter-specific
47
48
49 488 pairs with other goose species wintering in the same area, especially if there is a shortage of
50
51 489 conspecific mates. The bean goose has previously been reported to hybridize at least with the
52
53
54 490 pink-footed and the greater white fronted goose (McCarthy 2006; Kampe-Persson and Lerner
55
56 491 2007). These species use the same wintering areas as the bean goose, which may promote
57
58 492 interspecific pairing and hybridization. The locations from where the putative hybrids were
59
60
61
62
63
64
65

1
2 494
3
4
5 495
6
7 496
8
9
10 497
11
12 498
13
14 499
15
16
17 500
18
19 501
20
21
22 502
23
24 503
25
26
27 504
28
29 505
30
31 506
32
33
34 507
35
36 508
37
38
39 509
40
41 510
42
43
44 511
45
46 512
47
48
49 513
50
51 514
52

hunted are in concordance with the possible hybrid origin. The putative pink-footed goose x bean goose hybrid was hunted at the Finnish coast of the Bothnian Bay that is along the migration route of the pink-footed goose (Hölttä 2013) and the putative greater white-fronted goose x bean goose hybrid was hunted in south-eastern Finland along the migration route of the greater white-fronted goose (Mooij et al. 1999; Fig. 3).

53 515 **Taxonomy**

54
55
56 516
57
58 517
59
60
61
62
63
64
65

The taxonomy of the bean goose-pink-footed complex is still not completely resolved. Our phylogeny corresponds to that of Ruokonen et al. (2008) as the same mtDNA region was used. However, Ottenburghs et al. (2016a) show incongruence in phylogeny when different genetic markers are used. This incongruence is probably caused by incomplete lineage sorting or speciation with hybridization (Ruokonen et al. 2000; Ottenburghs et al. 2016a). Our results show that the cross-species microsatellite panel did not help to resolve the bean goose phylogeny. However, a new microsatellite panel developed for the bean goose (Kleven et al. 2016) could resolve the shortcomings of our panel and should be tested in further studies. Further, a thorough genomic analysis with sampling across the entire range of the bean goose (Ottenburghs et al. 2016a) with all the subspecies and closely related species involved, should be carried out in order to resolve the taxonomic relationship of the bean goose-pink-footed goose complex. Also, elucidating the pattern of incomplete lineage sorting could be useful in order to explore the evolutionary forces that have acted during speciation within the genus *Anser* as was done by Scally et al. (2012) with human-great ape whole-genome sequences.

53 515 **Management implications**

54
55
56 516
57
58 517
59
60
61
62
63
64
65

In this study we found that over half of the Finnish bean goose bag consists of the declining taiga bean geese and that the tundra bean goose portion of the bag comes mainly from south-eastern

1
2 518 Finland. Our estimate is that, on average, 2200 taiga bean geese per year were hunted in Finland
3
4 519 alone during our study period, which is far too many considering the fast decline of this
5
6 520 subspecies. On the contrary, hunting of the tundra bean goose with a large and stable population
7
8 521 could be permitted as long as it does not affect the taiga bean goose population. There is no
9
10 522 knowledge of exact cause for the decline of the taiga bean goose, but potential reasons could be
11
12 523 hunting (especially reproducing individuals), habitat destruction, increased predation, human
13
14 524 disturbance and climate change. Interspecific competition with increasing numbers of whooper
15
16 525 swans (*Cygnus cygnus*) at nesting sites (Kampe-Persson et al. 2005) or with other geese species
17
18 526 in staging and wintering sites have also been suggested but not proven in any studies.

19
20
21
22 527 Conservation actions have already been made in Finland. Hunting of bean geese was
23
24 528 seasonally restricted during 2010-2013, banned completely in 2014-2016 and a draft national
25
26 529 management plan to protect the taiga bean goose was produced in 2014 (The Finnish Ministry of
27
28 530 Agriculture and Forestry 2014). Also the International Taiga Bean Goose Management Plan was
29
30 531 published in 2015 (Marjakangas et al. 2015). This is the first flyway conservation plan for a
31
32 532 declining species that is still open for hunting. Unfortunately, genetic issues are not implemented
33
34 533 in either of these management plans. Further conservation actions should be made, including a
35
36 534 thorough study of the spatial population genetic structure of the breeding geese, continuation of
37
38 535 restrictions for hunting at the sites where *A. f. fabalis* is the most common subspecies (at least in
39
40 536 central and northern Finland) and management of breeding habitats.

41
42
43
44
45
46 537

47
48
49 538

50
51 539

52
53 540

54
55
56
57
58
59
60
61
62
63
64
65

541 **Literature cited**

- 1
2 542 Aarvak T, Øien IJ (2009) Monitoring of Bean Goose in Finnmark County, Norway –Results
3
4
5 543 from 2008. Norsk Ornitologisk Forening. NOF rapport:2-2009.
6
7 544 Allendorf FW, England PR, Luikart G, Ritchie PA, Ryman N (2008) Genetic effects of harvest
8
9
10 545 on wild animal populations. Trends Ecol Evol 23:327-37.
11
12 546 Avise JC, Alisauskas RT, Nelson WS, Ankney CD (1992) Matriarchal population genetic
13
14 547 structure in an avian species with female natal philopatry. Evolution 46:1084-96.
15
16
17 548 Bandelt H-J, Forster P, Röhl A (1999) Median-joining networks for inferring intraspecific
18
19 549 phylogenies. Mol Biol Evol 16:37-48.
20
21
22 550 Belkhir K, Borsa P, Chikhi L, Raufaste N, Bonhomme F (1996-2004) GENETIX 4.05, logiciel
23
24 551 sous Windows TM pour la génétique des populations. Laboratoire Génome, Populations,
25
26 552 Interactions, CNRS UMR 5171, Université de Montpellier II, Montpellier (France).
27
28
29 553 BirdLife International and NatureServe (2014) Bird Species Distribution Maps of the World.
30
31 554 2015. *Anser fabalis*. The IUCN Red List of Threatened Species. Version 2015-3.
32
33
34 555 <http://maps.iucnredlist.org/map.html?id=22679875>. Accessed 11 February 2016.
35
36 556 Burgers J, Smit JJ, Van Der Voet H (1991) Origins and systematics of two types of the bean
37
38
39 557 goose *Anser fabalis* (Latham, 1787) wintering in the Netherlands. Ardea 79:307-16.
40
41 558 Canada Goose Committee; Atlantic Flyway Council Game Bird Technical Section (2008) A
42
43 559 Management Plan for the Atlantic Population of Canada Geese.
44
45
46 560 http://dnr2.maryland.gov/wildlife/Documents/2008_CAGO_AP_MgtPlan.pdf. Accessed 3
47
48 561 November 2015.
49
50
51 562 Chapuis M, Estoup A (2007) Microsatellite null alleles and estimation of population
52
53 563 differentiation. Mol Biol Evol 24:621-31.
54
55
56 564 Council of Europe (2007) Convention on the Conservation of European Wildlife and Natural
57
58 565 Habitats. European Charter on Hunting and Biodiversity.
59
60
61
62
63
64
65

566 <https://wcd.coe.int/com.instranet.InstraServlet?command=com.instranet.CmdBlobGet&InstranetImage=1883368&SecMode=1&DocId=1436274&Usage=2>. Accessed 20 November
567
568 2015.

569 Cranswick PA, Raducescu L, Hilton GM, Petkov N (2012) International Single Species Action
570 Plan for the Conservation of the Red-breasted Goose (*Branta ruficollis*). AEWA Technical
571 Series No. 46.

572 De Jong A, Heinicke T, Aarvak T, Øien IJ (2013) Movements of tundra bean goose *Anser*
573 *fabalis rossicus* neck-banded in northern Scandinavia. *Ornis Norv* 36:28-31.

574 Delacour J (1951) Taxonomic notes on the bean geese, *Anser fabalis* Lath. *Ardea* 39:135-42.

575 Earl DA, vonHoldt BM (2012) STRUCTURE HARVESTER: a website and program for
576 visualizing STRUCTURE output and implementing the Evanno method. *Conserv Genet*
577 *Res* 4:359-61.

578 Evanno G, Regnaut S, Goudet J (2005) Detecting the number of clusters of individuals using the
579 software STRUCTURE: a simulation study. *Mol Ecol* 14:2611-20.

580 Excoffier L, Lischer HEL (2010) Arlequin suite ver 3.5: a new series of programs to perform
581 population genetics analyses under Linux and Windows. *Mol Ecol Res* 10:564-7.

582 Excoffier L, Smouse PE, Quattro JM (1992) Analysis of molecular variance inferred from metric
583 distances among DNA haplotypes: application to human mitochondrial DNA restriction
584 data. *Genetics* 131:479-91.

585 Falush D, Stephens M, Pritchard JK (2003) Inference of population structure using multilocus
586 genotype data: linked loci and correlated allele frequencies. *Genetics* 164:1567-87.

587 Fu Y-X (1997) Statistical tests of neutrality of mutations against population growth, hitchhiking
588 and background selection. *Genetics* 147:915-25.

589 Finnish Meteorological Institute (2015) Open data - Finnish Meteorological Institute.
590 <https://en.ilmatieteenlaitos.fi/open-data>. Accessed 2 December 2015.

- 591 Finnish Ministry of Agriculture and Forestry (2014) Suomen metsähanhikannan
1
2 592 hoitosuunnitelma; luonnos (in Finnish) [Finnish bean goose management plan; draft].
3
4 593 <http://docplayer.fi/1154252-Suomen-metsahanhikannan-hoitosuunnitelma.html>. Accessed
5
6
7 594 2 November 2015.
8
- 9
10 595 Fox AD, Ebbsing BS, Mitchell C, Heinicke T, Aarvak T, Colhoun K, Clausen P, Dereliev S,
11
12 596 Faragö S, Koffijberg K, Kruckenberg H, Loonen MJJE, Madsen J, Mooij J, Musil P,
13
14 597 Nilsson L, Pihl S, Van Der Jeugd H (2010) Current estimates of goose population sizes in
15
16 598 western Europe, a gap analysis and an assessment of trends. *Ornis Svec* 20:115-27.
17
18
- 19 599 Frankham R (2005) Genetics and extinction. *Biol Conserv* 126:131-40.
20
- 21
22 600 Frankham R (2010) Challenges and opportunities of genetic approaches to biological
23
24 601 conservation. *Biol Conserv* 143:1919-27.
25
- 26 602 Gese EM, Knowlton FF, Adams JR, Beck K, Fuller TK, Murrey DL, Steury TD, Stoskopf MK,
27
28 603 Waddell WT, Waits LP (2015) Managing hybridization of a recovering endangered
29
30 604 species: the red wolf *Canis rufus* as a case study. *Curr Zool* 61:191-205.
31
32
33
- 34 605 Goudet J (1995) FSTAT (Version 1.2): a computer program to calculate F-statistics. *J Hered*
35
36 606 86:485-6.
37
- 38
39 607 Grant PR, Grant BR (1992) Hybridization of bird species. *Science* 256:193-7.
40
- 41 608 Grant WS, Bowen BW (1998) Shallow population histories in deep evolutionary lineages of
42
43 609 marine fishes: insights from sardines and anchovies and lessons for conservation. *J Hered*
44
45 610 89:415-26.
46
47
- 48 611 Green M (2001) Is wind drift in migrating barnacle and brent geese, *Branta leucopsis* and *Branta*
49
50 612 *bernicla*, adaptive or non-adaptive? *Behav Ecol Sociobiol* 50:45-54.
51
52
- 53 613 Hall T (1999) BioEdit: a user-friendly biological sequence alignment editor and analysis
54
55 614 program for Windows 95/98/NT. *Nucleic Acids Symp Ser* 41:95-8.
56
57
58
59
60
61
62
63
64
65

- 1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65
- 615 Hardy OJ, Charbonnel N, Fréville H, Heuertz M (2003) Microsatellite allele sizes: a simple test
616 to assess their significance on genetic differentiation. *Genetics* 163:1467-82.
- 617 Hardy OJ, Vekemans X (2002) SPAGeDI: a versatile computer program to analyse spatial
618 genetic structure at the individual or population levels. *Mol Ecol Notes* 2:618-20.
- 619 Hasegawa M, Kishino H, Yano T (1985) Dating of the human-ape splitting by a molecular clock
620 of mitochondrial DNA. *J Mol Evol* 22:160-74.
- 621 Hedrick PW, Allendorf FW, Baker CS (2013) Estimation of male gene flow from measures of
622 nuclear and female genetic differentiation. *J Hered* 104:713-7.
- 623 Hedrick PW, Singh S, Aspi J (2015) Estimation of male gene flow: use caution. *J Hered*
624 106:745-8.
- 625 Hirschfeld A, Heyd A (2005) Mortality of migratory birds caused by hunting in Europe: bag
626 statistics and proposals for the conservation of birds and animal welfare. *Berichte*
627 *Vogelschutz* 42:47-74.
- 628 Hoffman JI, Amos W (2005) Microsatellite genotyping errors: detection approaches, common
629 sources and consequences for paternal exclusion. *Mol Ecol* 14:599-612.
- 630 Hölttä H (2013) Lintujen muuttoreitit ja pullonkaula-alueet Pohjois-Pohjanmaalla
631 tuulivoimarakentamisen kannalta (in Finnish) [Migration routes and bottleneck areas of
632 birds in Northern Ostrobothnia in relation to wind power construction]. BirdLife Finland.
633 http://www.birdlife.fi/suojelu/paikat/maali/Muuttoreittiselvitys_ja_liitteet_15032013.pdf.
634 Accessed 9 September 2015.
- 635 Inman RL, Scribner KT, Prince HH, Warrillow JA, Luukkonen DR, Padding PI (2003) A novel
636 method for Canada goose harvest derivation using genetic analysis of tail feathers. *Wildl*
637 *Soc Bull* 31:1126-31.

- 638 Johnson WE, Onorato DP, Roelke ME, Land ED, Cunningham M, Belden RC, McBride R,
1
2 639 Jansen D, Lotz M, Shindle D, Howard J, Wildt DE, Penfold LM, Hostetler JA, Oli MK,
3
4 640 O'Brien SJ (2010) Genetic restoration of the Florida panther. *Science* 329:1641-5.
5
6
7 641 Jones T, Martin K, Barov B, Nagy S (Compilers) (2008) International Single Species Action
8
9 642 Plan for the Conservation of the Western Palearctic Population of the Lesser White-fronted
10
11 643 Goose *Anser erythropus*. AEWA Technical Series No.36. Bonn, Germany.
12
13
14 644 Kampe-Persson H, Bildström L, Bildström M (2005) Can nesting competition with whooper
15
16 645 swan *Cygnus cygnus* cause a decline of the Swedish taiga goose *Anser fabalis fabalis*
17
18 646 population? (in Swedish with English summary) *Ornis Svec* 15:119-21.
19
20
21 647 Kampe-Persson H, Lerner H (2007) Occurrence of hybrid geese in Sweden - a conservation
22
23 648 problem? *Ornis Svec* 17:154-86.
24
25
26 649 Kleven O, Kroglund RT, Østnes JE (2016) Isolation, characterization and multiplex PCR
27
28 650 development of bean goose (*Anser fabalis*) microsatellite loci. *J Ornithol* 157:641-6.
29
30
31 651 Laikre L (2010) Genetic diversity is overlooked in international conservation policy
32
33 652 implementation. *Conserv Genet* 11:349-54.
34
35
36 653 Leigh JW, Bryant D (2015) POPART: full-feature software for haplotype network construction.
37
38 654 *Methods Ecol Evol* 6:1110-6.
39
40
41 655 Librado P, Rozas J (2009) DnaSP v5: a software for comprehensive analysis of DNA
42
43 656 polymorphism data. *Bioinformatics* 25:1451-2.
44
45
46 657 Lopez JV, Yuhki N, Masuda R, Modi W, O'Brien J (1994) Numt, a recent transfer and tandem
47
48 658 amplification of mitochondrial DNA to the nuclear genome of the domestic cat. *J Mol Evol*
49
50 659 39:174-90.
51
52
53 660 Madsen J (1991) Status and trends of goose populations in the western Palearctic in the 1980s.
54
55 661 *Ardea* 79:113-371.
56
57
58
59
60
61
62
63
64
65

- 662 Madsen J, Williams JH, (Compilers) (2012) International Species Management Plan for the
1
2 663 Svalbard Population of the Pink-footed Goose *Anser brachyrhynchus*. AEWA Technical
3
4 664 Series No.48. Bonn, Germany.
5
6
7 665 Marjakangas A, Alhainen M, Fox AD, Heinicke T, Madsen J, Nilsson L, Rozenfeld S
8
9 666 (Compilers) (2015) International Single Species Action Plan for the Conservation of the
10
11 667 Taiga Bean Goose *Anser fabalis fabalis* AEWA Technical Series No. 56, Bonn, Germany.
12
13
14 668 McCarthy EM (2006) Handbook of the Avian Hybrids of the World. Oxford University Press,
15
16 669 New York.
17
18
19 670 Mooij JH, Faragó S, Kirby JS (1999) White-fronted Goose *Anser albifrons albifrons*. In: Madsen
20
21 671 J, Cracknell G, Fox T (eds) Goose Populations in the Western Palearctic. A Review of
22
23 672 Status and Distribution. Wetlands International Publication No. 48. Wetlands International,
24
25 673 Wageningen, National Environmental Research Institute, Rønde, pp 94-129.
26
27
28
29 674 Moyle LC, Stinchcombe JR, Hudgens BR, Morris WF (2003) Conservation genetics in the
30
31 675 recovery of endangered animal species: a review of US endangered species recovery plans
32
33 676 (1977-1998). *Anim Biodiversity Conserv* 26:85-95.
34
35
36 677 Mylecraine KA, Gibbs HL, Anderson CS, Shieldcastle MC (2008) Using 2 genetic markers to
37
38 678 discriminate among Canada goose populations in Ohio. *J Wildl Manage* 72:1220-30.
39
40
41 679 Nilsson L (2011) The migrations of Finnish bean geese *Anser fabalis* in 1978-2011. *Ornis Svec*
42
43 680 21:157-66.
44
45
46 681 Nilsson L, Persson H (2001) Natal and breeding dispersal in the Baltic greylag goose *Anser*
47
48 682 *anser*. *Wildfowl* 52:21-30.
49
50
51 683 Nilsson L, van den Bergh L, Madsen J (1999) Taiga Bean Goose *Anser fabalis fabalis*. In:
52
53 684 Madsen J, Cracknell G, Fox T (eds) Goose Populations in the Western Palearctic. A
54
55 685 Review of Status and Distribution Wetlands International Publication No. 48. Wetlands
56
57 686 International, Wageningen, National Environmental Research Institute, Rønde, pp 23-39.
58
59
60
61
62
63
64
65

687 Noreikiene K, Teacher AGF, Madsen J, Gienapp P (2012) Isolation and characterization of 55
1
2 688 novel microsatellite markers for the pink-footed goose (*Anser brachyrhynchus*). Conserv
3
4 689 Genet Res 4:423-428.
5
6
7 690 Ottenburghs J, Megens H-, Kraus RHS, Madsen O, van Hooft P, van Wieren SE, Crooijmans
8
9 691 RPMA, Ydenberg RC, Groenen MAM, Prins HHT (2016a) A tree of geese: a
10
11 692 phylogenomic perspective on the evolutionary history of true geese. Mol Phylogenet Evol
12
13 693 101:303-13.
14
15
16 694 Ottenburghs J, van Hooft P, van Wieren SE, Ydenberg RC, Prins HHT (2016b) Hybridization in
17
18 695 geese: a review. Front Zool 13:20.
19
20
21 696 Pacific Flyway Council (2015) Pacific Flyway Management Plan for the Dusky Canada Goose.
22
23 697 Care of the U.S. Fish and Wildlife Service's Pacific Flyway Representative, Vancouver,
24
25 698 Washington. 41 pp.+ appendices. http://pacificflyway.gov/Documents/Dcg_plan.pdf.
26
27 699 Accessed 2 November 2015.
28
29
30
31 700 Palsbøll PJ, Bérubé M, Allendorf FW (2007) Identification of management units using
32
33 701 population genetic data. Trends Ecol Evol 22:11-6.
34
35
36 702 Pritchard JK, Stephens M, Donnelly P (2000) Inference of population structure using multilocus
37
38 703 genotype data. Genetics 155:945-59.
39
40
41 704 Rambaut A (2006-2014) FigTree, Tree Figure Drawing Tool, Version 1.4., available:
42
43 705 <http://tree.bio.ed.ac.uk/>.
44
45
46 706 Rambaut A, Suchard MA, Xie D, Drummond AJ (2014) Tracer v1.6, available from
47
48 707 <http://beast.bio.ed.ac.uk/Tracer>.
49
50
51 708 Ramos-Onsins SE, Rozas J (2002) Statistical properties of new neutrality tests against population
52
53 709 growth. Mol Biol Evol 19:2092-100.
54
55
56 710 Rice WR (1989) Analyzing tables of statistical tests. Evolution 43:223-5.
57
58
59
60
61
62
63
64
65

- 1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65
- 711 Rivers MC, Brummitt NA, NicLughadha E, Meagher TR (2014) Do species conservation
712 assessments capture genetic diversity? *Glob Ecol Conserv* 2:81-7.
- 713 Robinson JA, Colhoun K (Compilers) (2006) International Single Species Action Plan for the
714 Conservation of the Light-bellied Brent Goose (East Canadian High Arctic Population)
715 *Branta bernicla hrota*. AEWA Technical Series No. 11. Bonn, Germany.
- 716 Rohwer FC, Anderson MG (1988) Female-biased philopatry, monogamy, and the timing of pair
717 formation in migratory waterfowl. *Curr Ornithol* 5:187-221.
- 718 Ronquist F, Teslenko M, Van Der Mark P, Ayres DL, Darling A, Höhna S, Larget B, Liu L,
719 Suchard MA, Huelsenbeck JP (2012) MrBayes 3.2: efficient Bayesian phylogenetic
720 inference and model choice across a large model space. *Syst Biol* 61:539-42.
- 721 Rousset F (2008) GENEPOP'007: a complete re-implementation of the GENEPOP software for
722 Windows and Linux. *Mol Ecol Res* 8:103-6.
- 723 Ruokonen M, Aarvak T (2011) Typology revisited: historical taxa of the bean goose - pink-
724 footed goose complex. *Ardea* 99:103-12.
- 725 Ruokonen M, Aarvak T, Chesser RK, Lundqvist A-C, Merilä J (2010) Temporal increase in
726 mtDNA diversity in a declining population. *Mol Ecol* 19:2408-17.
- 727 Ruokonen M, Aarvak T, Madsen J (2005) Colonization history of the high-arctic pink-footed
728 goose *Anser brachyrhynchus*. *Mol Ecol* 14:171-8.
- 729 Ruokonen M, Andersson A-C, Tegelström H (2007) Using historical captive stocks in
730 conservation. The case of the lesser white-fronted goose. *Conserv Genet* 8:197-207.
- 731 Ruokonen M, Kvist L, Aarvak T, Markkola J, Morozov VV, Øien IJ, Syroechkovsky Jr. EE,
732 Tolvanen P, Lumme J (2004) Population genetic structure and conservation of the lesser
733 white-fronted goose *Anser erythropus*. *Conserv Genet* 5:501-12.
- 734 Ruokonen M, Kvist L, Lumme J (2000) Close relatedness between mitochondrial DNA from
735 seven *Anser* goose species. *J Evol Biol* 13:532-40.

- 736 Ruokonen M, Litvin K, Aarvak T (2008) Taxonomy of the bean goose-pink-footed goose. *Mol*
1
2 737 *Phylogenet Evol* 48:554-62.
3
4 738 Sangster G, Oreel GJ (1996) Progress in taxonomy of taiga and tundra bean geese. *Dutch*
5
6 *Birding* 18:310-6.
7 739
8
9 740 Scally A, Dutheil JY, Hillier LW, Jordan GE, Goodhead I, Herrero J, Hobolth A, Lappalainen T,
10
11 741 Mailund T, Marques-Bonet T, McCarthy S, Montgomery SH, Schwalie PC, Tang YA,
12
13 742 Ward MC, Xue Y, Yngvadottir B, Alkan C, Andersen LN, Ayub Q, Ball EV, Beal K,
14
15 743 Bradley BJ, Chen Y, Clee CM, Fitzgerald S, Graves TA, Gu Y, Heath P, Heger A,
16
17 744 Karakoc E, Kolb-Kokocinski A, Laird GK, Lunter G, Meader S, Mort M, Mullikin JC,
18
19 745 Munch K, O'Connor TD, Phillips AD, Prado-Martinez J, Rogers AS, Sajjadian S, Schmidt
20
21 746 D, Shaw K, Simpson JT, Stenson PD, Turner DJ, Vigilant L, Vilella AJ, Whitener W, Zhu
22
23 747 B, Cooper DN, De Jong P, Dermitzakis ET, Eichler EE, Flicek P, Goldman N, Mundy NI,
24
25 748 Ning Z, Odom DT, Ponting CP, Quail MA, Ryder OA, Searle SM, Warren WC, Wilson
26
27 749 RK, Schierup MH, Rogers J, Tyler-Smith C, Durbin R (2012) Insights into hominid
28
29 evolution from the gorilla genome sequence. *Nature* 483:169-75.
30
31 750
32
33 751 Scribner KT, Warrillow JA, Leafloor JO, Prince HH, Inman RL, Luukkonen DR, Flegel CS
34
35 (2003) Genetic methods for determining racial composition of Canada goose harvests. *J*
36
37 752 *Wildl Manage* 67:122-35.
38
39 753
40
41 754 Shorey RI, Scribner KT, Prince HH, Kravchenko AN, Luukkonen DR, Padding PI (2007)
42
43 755 Genetic analysis of standardized collections of cackling and Canada goose harvests. *J*
44
45 756 *Wildl Manage* 71:1458-66.
46
47
48 757 Short LL (1969) Taxonomic aspects of avian hybridization. *Auk* 86:84-105.
49
50
51 758 Slatkin M (1995) A measure of population subdivision based on microsatellite allele frequencies.
52
53 759 *Genetics* 139:457-62.
54
55
56
57
58
59
60
61
62
63
64
65

- 760 Stroud DA, Fox AD, Urquhart C, Francis IS (Compilers) (2012) International Single Species
1
2 761 Action Plan for the Conservation of the Greenland White-fronted Goose (*Anser albifrons*
3
4 762 *flavirostris*). AEWA Technical Series No. 45. Bonn, Germany.
5
6
7 763 Tajima F (1989) Statistical method for testing the neutral mutation hypothesis by DNA
8
9 764 polymorphism. *Genetics* 123:585-95.
10
11 765 Tamura K, Nei M (1993) Estimation of the number of nucleotide substitutions in the control
12
13 766 region of mitochondrial DNA in humans and chimpanzees. *Mol Biol Evol* 10:512-26.
14
15 767 Tamura K, Stecher G, Peterson D, Filipinski A, Kumar S (2013) MEGA6: molecular evolutionary
16
17 768 genetics analysis version 6.0. *Mol Biol Evol* 30:2725-9.
18
19 769 Todesco M, Pascual MA, Owens GL, Ostevik KL, Moyers BT, Hübner S, Heredia SM, Hahn
20
21 770 MA, Caseys C, Bock DG, Rieseberg LH (2016) Hybridization and extinction. *Evol Appl*
22
23 771 9:892-908.
24
25 772 Toivainen T, Metsänen T, Lehtiniemi T (2014) Lintujen päämuuttoreitit Suomessa (in Finnish)
26
27 773 [Main migration routes of birds in Finland]. BirdLife Finland.
28
29 774 U.S. Fish and Wildlife Service (2008) Florida Panther Recovery Plan (*Puma concolor coryi*),
30
31 775 Third Revision. U.S. Fish and Wildlife Service. Atlanta, Georgia.
32
33 776 http://ecos.fws.gov/docs/recovery_plan/081218.pdf. Accessed 12 February 2016.
34
35 777 U.S. Fish and Wildlife Service (2013) Red Wolf Adaptive Management Plan FY13-FY15.
36
37 778 http://www.fws.gov/redwolf/Images/20130211_RWAMP_2013-2015.pdf. Accessed 19
38
39 779 February 2016.
40
41 780 Valière N (2002) GIMLET: a computer program for analysing genetic individual identification
42
43 781 data. *Mol Ecol Notes* 2:377-9.
44
45 782 Van den Bergh L (1999) Tundra Bean Goose *Anser fabalis rossicus*. In: Madsen J, Cracknell G,
46
47 783 Fox T (eds) *Goose Populations in the Western Palearctic. A Review of Status and*
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

784 Distribution. Wetlands International Publication No. 48. Wetlands International,
1
2 785 Wageningen, National Environmental Research Institute, Rønde, pp 41-49.
3
4 786 Van Oosterhout C, Hutchinson WF, Wills DPM, Shipley P (2004) MICRO-CHECKER:
5
6 software for identifying and correcting genotyping errors in microsatellite data. Mol Ecol
7 787
8 Notes 4:535-8.
9 788
10
11 789 Waples RS, Gaggiotti O (2006) What is a population? An empirical evaluation of some genetic
12
13 methods for identifying the number of gene pools and their degree of connectivity. Mol
14 790
15 Ecol 15:1419-39.
16 791
17
18 792 Wetlands International (2016) Waterbird Population Estimates. <http://wpe.wetlands.org/>.
19
20 Accessed 11 January 2016.
21 793
22
23 794 Wright S (1951) The genetical structure of populations. Ann Eugen 15:323-54.
24
25 795 Zink RM, Barrowclough GF (2008) Mitochondrial DNA under siege in avian phylogeography.
26
27 Mol Ecol 17:2107-21.
28 796
29
30 797
31
32 798
33
34 799
35
36 800
37
38 801
39
40 802
41
42 803
43
44 804
45
46 805
47
48 806
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

807 **Figure captions**

1
2 808 **Fig. 1** Approximate breeding ranges of the bean goose subspecies (*Anser fabalis fabalis*, *A. f.*
3
4 809 *rossicus*, *A. f. serrirostris* and *A. f. middendorffii*) marked with the subspecies name and their
5
6
7 810 wintering ranges marked with dark shading (redrawn from BirdLife International and
8
9
10 811 NatureServe 2014 and from Ruokonen et al. 2008). The western subspecies *A. f. fabalis* and *A. f.*
11
12 812 *rossicus* overwinter in Europe and some eastern *A. f. fabalis* in central Asia. The eastern
13
14 813 subspecies *A. f. serrirostris* and *A. f. middendorffii* overwinter in China, Korea and Japan. The
15
16
17 814 sampling locations of the samples with known origin are marked with closed circles; for the
18
19 815 locations of the hunted geese, see Fig. 3

20
21
22 816 **Fig. 2** Median-joining network of the bean goose (*Anser fabalis*) haplotypes and outgroups (*A.*
23
24 817 *brachyrhynchus*, *A. albifrons*) for the mitochondrial control region (1235 bp). The haplotypes
25
26
27 818 named with three uppercase letters were previously described in Ruokonen et al. (2008). Letters
28
29 819 a or b in the haplotype names denote haplotypes that were identical with the 219 bp
30
31 820 hypervariable region but differ in the whole control region. The size of each circle is
32
33
34 821 proportional to the frequency of each haplotype. Black slashes across branches indicate the
35
36 822 number of mutational changes between the haplotypes

37
38
39 823 **Fig. 3** Geographical distribution of the harvest locations of the Finnish bean goose subspecies
40
41 824 (*Anser fabalis fabalis*, *A. f. rossicus* and *A. f. serrirostris*) based on mtDNA together with cases
42
43 825 of mtDNA introgression, i.e. where the bean goose had the mtDNA of a different species, either
44
45
46 826 *A. brachyrhynchus* or *A. albifrons*. The size of each circle is proportional to the frequency of
47
48
49 827 each subspecies

50
51 828 **Fig. 4** Microsatellite Factorial Correspondence Analysis (FCA) plot for the bean goose (*Anser*
52
53 829 *fabalis*). The subspecies (*A. f. fabalis*, *A. f. rossicus*, *A. f. serrirostris* and *A. f. middendorffii*)
54
55
56 830 were assigned based on mtDNA. The suggested hybrids with introgressed mtDNA (*A.*
57
58 831 *brachyrhynchus* and *A. albifrons*) are also indicated

Dear Editor-in-Chief, Christian Gortázar

We are pleased to send you our revised manuscript (Ref.: Ms. No. EJWR-D-16-00127R2 Determining the subspecies composition of bean goose harvests in Finland using genetic methods European Journal of Wildlife Research). We have followed all the suggestions made by the reviewer and we have also shortened the Discussion of our manuscript. We have also made additional minor changes to the manuscript when, through our reading, we have noticed lack of clarity. We believe that our manuscript has further improved and hope that it is now suited for publication. Below you can find our responses to the reviewer's comments marked in blue italics.

Sincerely yours,
Johanna Honka

Dear Dr Honka,
one of the reviewers who had evaluated the original submission of your manuscript has commented on your revised version. As you can see the manuscript is now mostly satisfying, yet some minor changes are pending. As the Discussion is still a bit long, I invite you to undertake a text reduction, for instance summarizing results of other comparable studies.

Best regards
Massimo Scandura

We have shortened the Discussion by condensing the text and summarising results of other studies (please see below).

Reviewer #1: I have re-reviewed the manuscript by Honka et al. and find that it has been significantly improved compared to the original submission. The authors have done a very good job thoroughly addressing the reviewers' comments. In my opinion, Discussion is still long and a heavy to be read. It takes about 35% of the total length, more or less. I have a series of minor remarks that I hope can help further the understanding of the MS.

We have now shortened our Discussion, as suggested, from 2658 words to 2046 words, which equals two pages, hopefully making it also easier to read now.

Line xx:

28- delete comma after second parenthesis, and change "facilitate management" with "aid conservation";
34- delete "From...," and use "We estimated"; delete "were made ...diversity,"
38- move " We also ...species." to the end of the abstract
54- "Among the three conventionally recognized levels of biodiversity", I mean, there are many indeed ...
55- delete "overlooked and"
57- delete "the mostspecies";
81- delete "and the harmful.... to" and replace with "as it has been proved in"

85- it could be nice spending just some words to explain differences between "hybridization" and "introgression", yet just in case authors agree to shorten a bit other parts in Discussion, otherwise Introduction too risks to be long as well;
91/92- delete "has a goal....and"
109- delete ", but the" and replace with ". However, taxonomy"
118/120: please, revise English
124- replace "as its own" with "separate"
128- Guess you meant "rapid speciation caused by hybridization", isn't it?
134- replace "environments" with "habitats"; delete "as shown" and use "(Figure 1)"
135- replace ", and" with ", whereas"
147- add "then" after "since"
154- ratio, in italic
158- delete "nuclear"
159- delete "as these ..challenging"
161-163: please, revise English and style
192- delete "to...amplification"
230/231- use capitals for spelling acronyms

We have now revised all the wording, English and style, except we did not change the word 'ratio', to italics because we actually couldn't see why it was suggested.

** 241- Why did you exclude gaps?

We performed maximum likelihood analyses both including gaps and excluding gaps and chose to present only the latter as the same tree topology was recovered, as now explained in the text.

255: Linkage Disequilibrium, LE; Hardy Weinberg Equilibrium, HWE
261-266: suggest removing paragraph
313-316: revise English and style
320- Mismatch .. add "Distribution (MD)"
322- R₂, with 2 as subscript
413/415- revise English (syntax)
418/421- as above

We have revised all the wording, English and style.

** 471-476: so, I understand that, up to you, high haplotype diversity is 0.68-0.86 whereas low haplotype diversity is 0.58: on which basis? Can you please test for these values to prove they are statistically different? "High" and "low" are qualitative descriptors only; as to me, 0.58 is not so low at all, depends on many things, you know.

We mean that 0.58 is low compared to other bean goose subspecies and species with high population numbers, we are sorry for being unclear. 0.58 is not as low as in the populations that are declining or have a historically low population size ($h=0.37-0.53$; lesser white fronted and pink-footed goose) but much less than species with a large population size such as the greater white fronted goose ($h=0.89$). We have now made efforts to make our comparisons clearer.

527-: "propensity to hybridization" (?)
525-529: see repetitions(hybrid/hybridization)
562: use "carried out" instead of "executed"
570: use "on average" between commas

575- may be a stop after "Finland" would be better

We have revised all the wording, English and style. We removed repetitions of hybridization by rewording.

Johanna Honka
Genetics and Physiology Unit
University of Oulu
Oulu 90014
Finland
johanna.honka@oulu.fi

Christian Gortázar
Chief Editor
European Journal of Wildlife Research

April 22, 2016

Dear Dr. Gortázar

We are pleased to submit an original research article “Genetic methods for determining subspecies composition of bean goose harvests in Finland” for consideration for publication in the European Journal of Wildlife Research.

We believe that this manuscript is appropriate for publication by the European Journal of Wildlife Research because our manuscript focuses on wildlife and conservation genetics with clear management implications. Our aim was to produce genetic data to aid management decisions of a declining taiga bean goose subspecies (*Anser fabalis fabalis*) that is hunted throughout Europe. This research is the first study that genetically quantifies the subspecies composition of the bean goose harvests. Our study showed that most of the harvested geese belong to the declining taiga bean goose subspecies and that the harvesting of the abundant tundra bean goose (*A. f. rossicus*) subspecies was geographically restricted to southeastern Finland. We also detected eastern tundra bean goose (*A. f. serrirostris*) haplotypes and possible cases of interspecific hybridization between goose species.

This manuscript has not been published and is not under consideration for publication elsewhere. We declare no conflict of interest.

Thank you for your consideration of our manuscript.

Sincerely,
Johanna Honka

Table 1 Population genetics statistics for the bean goose subspecies (*Anser fabalis fabalis*, *A. f. rossicus*, *A. f. serrirostris* and *A. f. middendorffii*). Above, data for mtDNA including sample size (n), number of haplotypes (H), haplotype diversities (h) and nucleotide diversities (π). Demographic population expansion tested with Tajima's D (D) and Fu's F_s (F_s), sum of squared deviation (SSD), Raggedness-index and R_2 statistics. SD stands for standard deviation. Middle, data for microsatellites including sample size (n), number of alleles (A), allelic richness (A_R), observed (H_O) and expected heterozygosity (H_E), unique alleles and Hardy-Weinberg equilibrium test (χ^2 , df=degrees of freedom). Below, data for microsatellite allele size ranges, unbiased probability of identity (P_{ID}) and probability of identity of siblings (P_{ID} Sib). Allele sequences and repeated motifs can be found in Noreikiene et al. (2012). Statistically significant values ($P < 0.05$) indicated with an asterisk and ($P < 0.001$) with two asterisks

Subspecies	n	H	h (SD)	π (SD)	D	F_s	SSD	Raggedness	R_2
<i>fabalis</i>	55	7	0.582 (0.062)	0.00103 (0.00011)	-0.182	-1.167	0.031	0.130	0.085
<i>rossicus</i>	49	9	0.676 (0.060)	0.00111 (0.00030)	-2.048*	-2.905	0.006	0.029*	0.068
<i>serrirostris</i>	4	3	0.733 (0.155)	0.00143 (0.00035)	1.386	0.688	0.108	0.293	0.278
<i>middendorffii</i>	9	6	0.855 (0.085)	0.00169 (0.00025)	-0.164	-1.607	0.383	0.044	0.140
	n	A	A_R	H_O	H_E	Unique alleles	χ^2	df	
<i>fabalis</i>	63	96	3.834	0.502	0.571	8	∞^{**}	32	
<i>rossicus</i>	68	91	3.690	0.490	0.566	7	∞^{**}	32	
<i>serrirostris</i>	8	54	3.290	0.430	0.510	0	34.81	26	
<i>middendorffii</i>	14	68	3.581	0.455	0.549	2	50.81	30	
Locus	Allele size range (bp)		P_{ID} (locus)	P_{ID} Sib (locus)					
Abra2	91-107		0.074	0.379					
Abra5	119-131		0.129	0.430					
Abra7	93-119		0.084	0.399					

Abra9	180-196	0.172	0.483
Abra10	147-169	0.088	0.394
Abra12	95-109	0.166	0.475
Abra14	150-154	0.752	0.871
Abra15	188-202	0.101	0.408
Abra19	168-184	0.105	0.408
Abra23	272-288	0.116	0.415
Abra24	289-293	0.475	0.696
Abra29	205-207	0.912	0.956
Abra30	108-120	0.315	0.576
Abra39	118-139	0.156	0.470
Abra43	120-134	0.203	0.502
Abra68	120-130	0.303	0.575

Table 2 Pairwise ϕ_{ST} values for bean goose subspecies (*Anser fabalis fabalis*, *A. f. rossicus*, *A. f. serrastris* and *A. f. middendorffii*) for mtDNA above the diagonal and the pairwise F_{ST} values and the R_{ST} values (in parentheses) for microsatellites below diagonal. Statistically significant (after Bonferroni correction) values ($P < 0.05$) indicated with an asterisk and ($P < 0.001$) with two asterisks

Subspecies	<i>fabalis</i>	<i>rossicus</i>	<i>serrastris</i>	<i>middendorffii</i>
<i>fabalis</i>		0.842**	0.791**	0.860**
<i>rossicus</i>	0.0256* (0.0058*)		0.679**	0.861**
<i>serrastris</i>	0.0314* (0.0323*)	0.0120 (0.0255)		0.780**
<i>middendorffii</i>	0.0096 (0.0251)	0.0271* (0.0526*)	0.0282 (0.0664*)	

Table 3 The percentage of the harvested Finnish bean geese subspecies assigned to subspecies by mtDNA (*Anser fabalis fabalis*, *A. f. rossicus* and *A. f. serrirostris*). Year 2013 was excluded due to a low sample size.

	2010	2011	2012
Sample size	9	24	62
<i>fabalis</i>	44.4	76.0	42.2
<i>rossicus</i>	55.6	20.0	54.7
<i>serrirostris</i>		4.0	1.6
mtDNA of other species			1.6

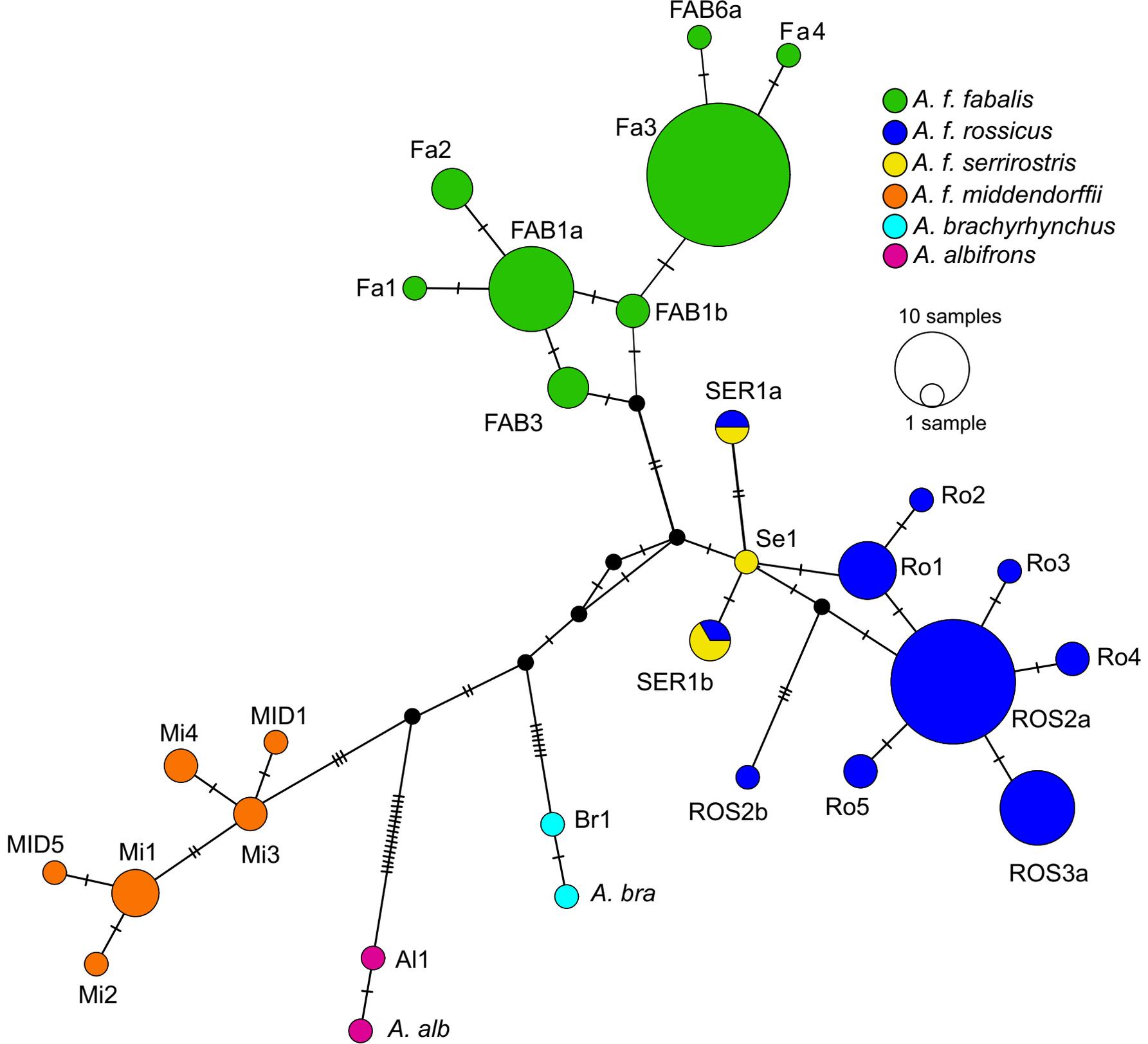
Table 4. F_{ST} values based on the microsatellites, differentiation in females based on the mtDNA ($F_{ST(f)}$) from ϕ_{ST} , calculated differentiation in males ($F_{ST(m)}$; Hedrick et al. 2013: equation 7a) and the ratio of the gene flow in males and females (m_m/m_f ; Hedrick et al. 2013: equation 7b) in bean goose subspecies (*Anser fabalis fabalis*, *A. f. rossicus*, *A. f. serrirostris* and *A. f. middendorffii*)

		F_{ST}	$F_{ST(f)}$	$F_{ST(m)}$	m_m/m_f
<i>fabalis</i>	<i>rossicus</i>	0.026	0.842	0.050	100.3
<i>fabalis</i>	<i>serrirostris</i>	0.031	0.791	0.062	57.24
<i>fabalis</i>	<i>middendorffii</i>	0.010	0.860	0.019	316.1
<i>rossicus</i>	<i>serrirostris</i>	0.012	0.679	0.023	89.03
<i>rossicus</i>	<i>middendorffii</i>	0.027	0.861	0.053	109.9
<i>serrirostris</i>	<i>middendorffii</i>	0.028	0.780	0.056	60.03

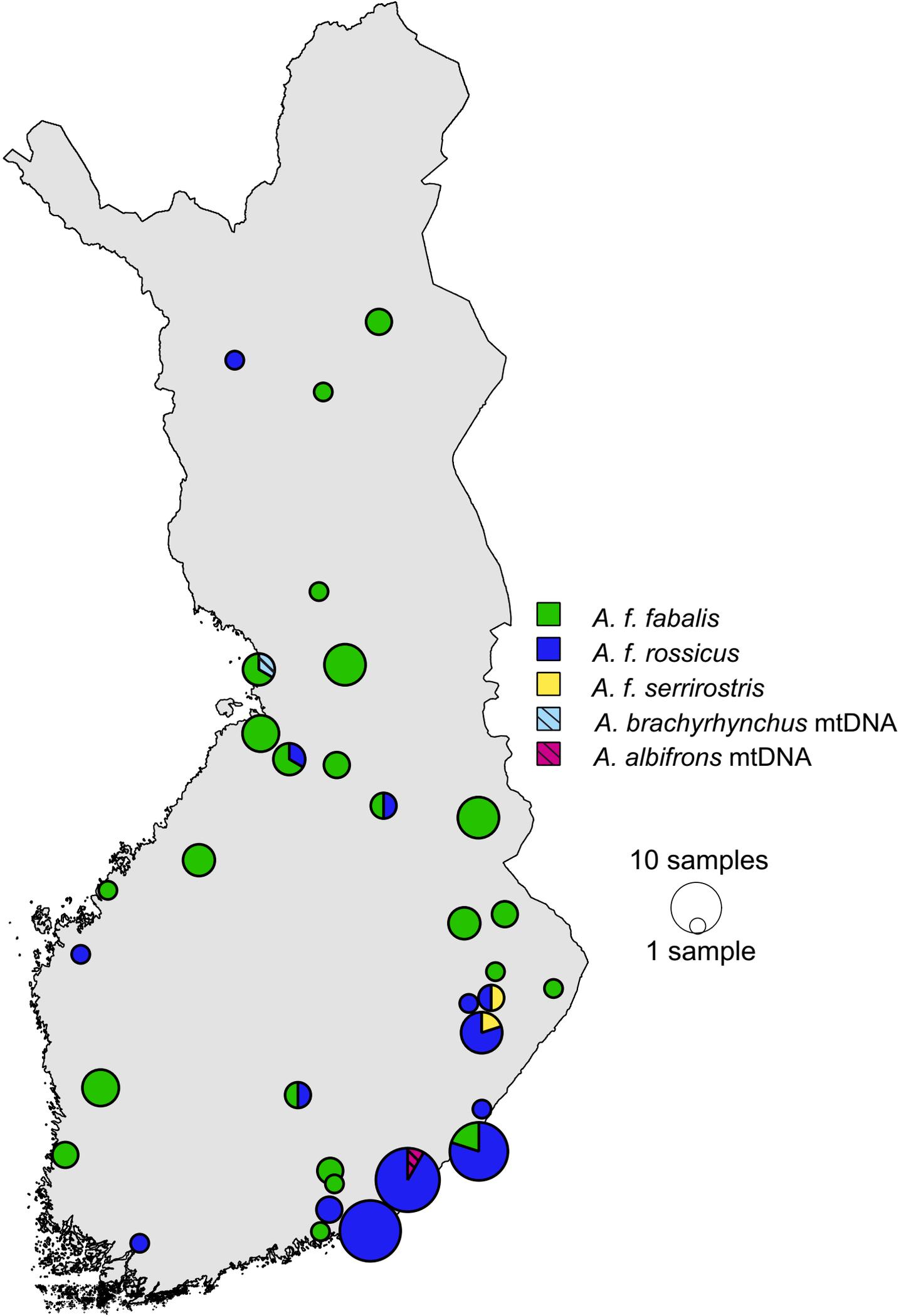
Figure



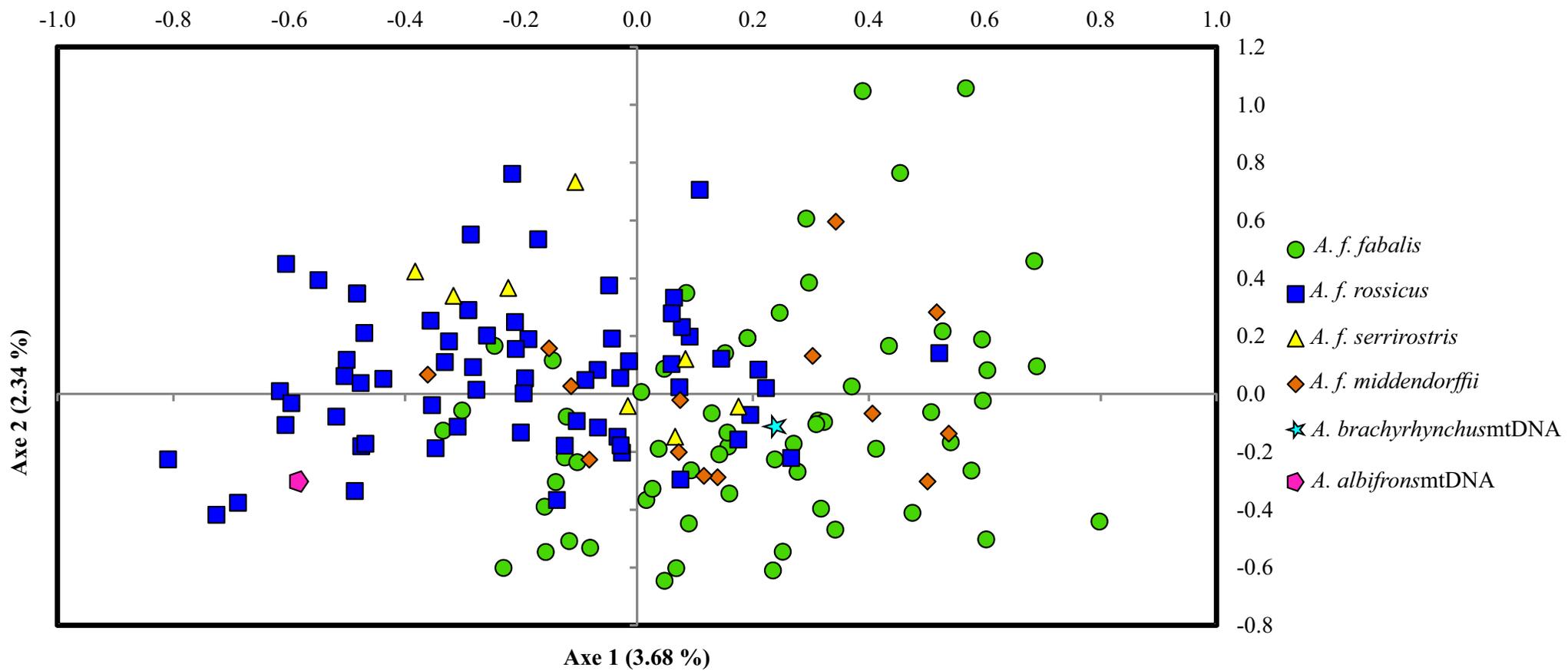
Figure



Figure

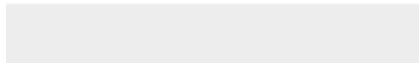
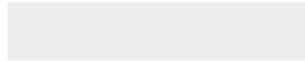


Figure





Click here to access/download
Electronic Supplementary Material
Online resource 1.doc



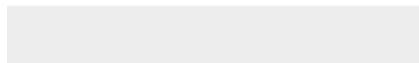


Click here to access/download
Electronic Supplementary Material
Online resource 2.doc





Click here to access/download
Electronic Supplementary Material
Online resource 3.doc





Click here to access/download
Electronic Supplementary Material
Online resource 4.doc

