1	Unravelling the mechanisms of PFOS toxicity by combining morphological and
2	transcriptomic analyses in zebrafish embryos
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### <u>Highlights</u>

- Zebrafish embryos show macroscopic alterations at PFOS concentrations > 1mg/L
- Transcriptomic changes occur at concentrations 1/10 to 1/100 the macroscopic LOEC
- Transcript analyses suggest alterations on lipid metabolism and the immune system
- Some of the observed changes occur at concentrations already detected in humans
- We propose tighter regulations limiting human and environmental exposure to PFOS

#### 25 Abstract

26 Exposure to PFOS (perfluorooctanesulfonate) has been related to toxic effects on lipid metabolism, immunological response, and different endocrine systems. We 27 28 present here a transcriptomic analysis of zebrafish embryos exposed to different 29 concentrations of PFOS (0.03-1.0 mg/L) from 48 to 120 hpf. No major survival or 30 morphological alterations (swimming bladder inflation, kyphosis, eye separation and 31 size...) were observed below the 1.0 mg/L mark. Conversely, we observed significant 32 increase in transcripts related to lipid transport and metabolism even at the lowest 33 used concentration. In addition, we observed a general decrease on transcripts related 34 to natural immunity and defense again infections, which adds to the recent concerns 35 about PFOS as immunotoxicant, particularly in humans. Derived PoD (Point of 36 Departure) values for transcriptional changes (0.011 mg/L) were about 200-fold lower 37 than the corresponding PoD values for morphometric effects (2.53 mg/L), and close to 38 levels observed in human blood serum or bird eggs. Our data suggest that currently 39 applicable tolerable levels of PFOS in commercial goods should be re-evaluated, taking 40 into account its potential effects on lipid metabolism and the immune system.

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Keywords: endocrine disrupting chemicals, high-throughput sequencing, differentially
expressed genes (DEGs), ANOVA-PLS, immune system, lipid disruption, benchmark
doses.

#### 46 1. Introduction

47 Polyfluoroalkyl and perfluoroalkyl substances (PFASs) consist of fully fluorinated 48 hydrophobic alkyl chains attached to different hydrophilic ending groups. Their unique 49 thermal stability and oxidation resistance have fueled their use for more than 50 years 50 in different industrial and domestic applications, as surfactants, surface treatment, fire 51 retardants, and coating materials (OECD, 2002). However, these same properties make 52 PFAS very bioaccumulative and persistent in the environment, with the capability for 53 long range transport through the atmosphere and water (Ahrens and Bundschuh, 54 2014). Perfluorooctane sulfonate (PFOS) is one of the most frequently detected PFASs 55 in the environment, and the first one to be added to Annex B of the Stockholm 56 Convention on Persistent Organic Pollutants list, which resulted in a global restriction, 57 but not a ban, on its production and use (Paul et al., 2009). Like other PFAS, exposure 58 to PFOS is known to induce adverse effects on growth, birth weight, fertility, 59 carcinogenesis, immunity, lipid metabolism, and the thyroid system (Chaparro-Ortega 60 et al., 2018; Du et al., 2013; Jensen and Leffers; Lau, 2015). Nevertheless, neither PFOS 61 nor any other PFAS has been yet categorized as endocrine disrupting chemical (EDC) by 62 the European Commission (European Commission, 2000) or by other regulatory 63 agencies. PFOS is found at relatively high levels (ng/L,  $\mu$ g/g and mg/L) in wastewaters, wastewater treatment plants' sludge, and exposed aquatic biota, respectively, showing 64 65 an extremely high bioaccumulation potential (Ahrens and Bundschuh, 2014; Arvaniti 66 and Stasinakis, 2015; Loos et al., 2013). In humans, blood serum levels of PFOS and 67 other PFAS ranged from 1 to 10  $\mu$ g/L in multiple surveys of general populations, and 68 from 100 to 1000  $\mu$ g/L for highly exposed populations (Kato et al., 2015; Olsen, 2015). 69 Toxicological studies revealed potential toxic effects of these substances ranging from growth and reproductive functions to lipid metabolism and oxidative stress (Ankley et 70 71 al., 2005; Lau et al., 2007; Rodriguez-Jorquera et al., 2019). The effect of the exposure 72 to PFOS and other related compounds on the immunologic system is a matter of 73 increasing concern, as it has been observed both in fish and in mammals, but its 74 significance for human health is largely unknown (Corsini et al., 2011; DeWitt et al., 75 2009; DeWitt et al., 2016). Moreover, a recent report on the effect of PFOS on mice 76 gut microbiome suggests an until now unexplored potentially harmful interactions 77 between PFOS and microbial and animal metabolisms (Lai et al., 2018).

79 In the present study, we aimed to characterize and to describe the mechanisms 80 of PFOS toxicity using zebrafish embryos as a model. Zebrafish characteristics (easy 81 maintenance, small size, short life cycle, embryo transparency, large offspring, etc.) 82 made it a preferred animal model for toxicity studies during the last decades (Hill et al., 83 2005; Scholz and Mayer, 2008; Stegeman et al., 2010). The study of zebrafish at least 84 at genomic, transcriptomic, proteomic and metabolomic levels (Mushtaq et al., 2013) 85 and its recognition as an acceptable vertebrate model for human and environmental 86 toxicology (Raldúa and Piña, 2014; Strahle et al., 2012), make it an excellent election 87 for ecotoxicological studies. During the last years, high-throughput next generation 88 sequencing (HT-NGS) technologies facilitated the improvement and achievements of 89 the transcriptomic studies (Mortazavi et al., 2008; Reuter et al., 2015), and showed its 90 usefulness in the study of transcriptomic effects of several toxicants (Baker and 91 Hardiman, 2014; Caballero-Gallardo et al., 2016). Although transcriptomics effects of 92 PFOS over zebrafish larvae have been previously assessed, these studies were 93 centered in a targeted dataset of transcripts (Jantzen et al., 2016; Shi et al., 2008) or 94 based in a single PFOS dose (Chen et al., 2014; Fai Tse et al., 2016). The aim of this 95 work was to analyze the toxic effects of PFOS in zebrafish embryos at sub-lethal 96 concentrations, using a combination of morphometric and transcriptomic techniques. 97 For that reason, we have designed dose-response assays using different PFOS 98 concentrations to unravel the different mechanisms underlying the different toxic 99 effects and to contribute to the risk assessment analysis of these currently ubiquitous 100 pollutants.

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#### 102 2. Materials and methods

#### 103 **2.1.** Zebrafish maintenance and rearing conditions

Adult wild-type zebrafish (*Danio rerio*, 12–18 months old) were maintained under controlled standard conditions ( $28 \pm 1 \,^{\circ}$ C, 12L:12D photoperiod,  $\leq 5 \,$  fish/L) in fish water. Fish water was composed of 90 µg/ml of Instant Ocean (Aquarium Systems, Sarrebourg, France) and 0.58 mM of CaSO<sub>4</sub>·2H<sub>2</sub>O, dissolved in reverse osmosis purified water. Zebrafish were fed twice a day with dry flakes (TetraMin, Tetra, Germany). Eggs were obtained by natural mating of the adults, placing in breeding tanks males and

110 females in a 2:1 proportion, respectively. A mesh was placed in each breeding tank to 111 avoid zebrafish access to eggs, which were collected and rinsed at 2 hpf (hours post 112 fertilization). Fertilization rate was assessed to be at least 70 % (OECD, 2013) and 113 fertilized eggs were randomly placed in 6 well plates with fish water (under standard 114 conditions) at a density of 3 embryos/mL. Fish water was changed daily until the start 115 of the PFOS exposure at 48 hpf (sections 2.2.2. and 2.2.3.). All procedures were 116 performed accordingly with the institutional guidelines under a license from the local 117 government (DAMM 7669, 7964) and were approved by the Institutional Animal Care 118 and Use Committees at the Research and Development Centre of the Spanish National 119 Research Council (CID-CSIC).

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#### 121 **2.2.** Zebrafish eleutheroembryo exposure to PFOS

#### 122 **2.2.1. PFOS solutions preparation.**

123 PFOS (perfluorooctanesulfonate (PFOS, CAS-RN: 2795-39-3) potassium salt) was purchased from Sigma-Aldrich (St. Louis, MO, USA, ≥98% purity). Five hundred-fold 124 125 stock solutions (50-5000 mg/L, depending on the experiment, see sections 2.2.2. and 126 2.3.3.) were prepared in dimethyl sulfoxide (DMSO) and stored at -20 °C. Experimental 127 solutions were prepared every day by dilution of the stock with fish water, with a final 128 DMSO concentration of 0.2% (v/v) in all exposed and control groups. The pH of the 129 working solutions was in the recommended range (6.8 - 7.5) (Avdesh et al., 2012). 130 Concentrations are given as nominal values. Since the stability of PFOS in water 131 solutions for at least 24h have been previously assessed (Kato et al., 2013; Lyu et al., 132 2015), daily water changes were considered sufficient to ensure constant PFOS 133 concentrations.

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#### 135 **2.2.2.** Exposures for morphometric tests

2ebrafish eleutheroembryos were exposed from 2 dpf to 5 dpf to a wide range of PFOS concentrations to stablish a suitable LOAEC (Lowest Observed Adverse Effect Concentration) below which transcriptomics studies should take place. We did not expose embryos during the first 48 hpf to avoid confounding factors due to the interference of PFOS with the early embryonic processes and to focus in the effect of PFOS in the already differentiated tissues of the larvae. PFOS concentrations ranged 142 from 0 (control, 0.2% of DMSO) to PFOS solutions containing 0.10, 0.25, 0.50, 1.0, 2.5, 143 5.0, 7.5, 10 and 100 mg/L of PFOS (all of them in the presence of 0.2% of DMSO). 144 Anatomical development of embryos was followed daily during the exposure as 145 described (Kimmel et al., 1995). Survival (3, 4 and 5 dpf), hatching (3, 4 and 5 dpf) and 146 swim bladder inflation rates (4 and 5 dpf) were assessed in at least 50 larvae per each 147 experimental group.

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#### **2.2.3.** Exposure for transcriptome analysis

150 Using the morphometric data, a second round of PFOS exposure was 151 performed, avoiding PFOS concentrations that could affect embryo viability. Therefore, 152 the highest used concentration was 1.0 mg/L of PFOS, as determined as LOAEC for 153 morphometric effects (see section 3.2.). Zebrafish eleutheroembryos were exposed to 154 control (0.2% DMSO), 0.03, 0.3 and 1.0 mg/L of PFOS during 72h from 2 to 5 dpf. The 155 experimental design was selected to minimize all possible covariates. All animals 156 belonged to the same batch and were simultaneously incubated in 6-well plates, each 157 one containing control and exposed groups for each concentration (no batch- or plate-158 bias). Fresh working solutions were daily prepared from the stocks and renewed as 159 previously described. For each experimental group, anatomical development of 160 embryos (Kimmel et al., 1995), sub-lethal and developmental efects, survival and 161 hatching rates at 3, 4 and 5 dpf, and swim bladder inflation rates at 4 and 5 dpf were 162 reported, according to the OECD 236 guidelines (OECD, 2013). Replicates of 10 163 eleutheroembryos per experimental condition were collected, snap-frozen in dry ice 164 and stored at -80 °C until further analysis (RNA extraction for high-throughput 165 sequencing and RT-qPCR confirmation purposes).

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#### 167 2. 3. Eleutheroembryo fixation, morphological and immunochemical measurements

168 Zebrafish eleutheroembryos were collected and fixed overnight at 4 °C in a 169 phosphate buffered saline 1X solution (PBS) with 4% PFA (paraformaldehyde) as 170 previously described (Martínez et al., 2018; Raldua et al., 2008). For morphological 171 measurements, fixed eleutheroembryos were then washed several times with PBS and 172 gradually transferred to 90% glycerol (10% PBS 1X) for long-term preservation and 173 positioning facilitation under the microscope. A stereomicroscope Nikon SMZ1500

174 equipped with a Nikon digital Sight DS-Ri1 camera was used to acquire lateral and 175 dorsoventral images of the fixed embryos (Figure 1A). Afterwards the following 176 morphological parameters were measured using the free graphical image analysis 177 software ImageJ (National Institutes of Health, Bethesda, MD, USA): Body length (BL), 178 head-trunk angle (HTA), yolk sac area (YSA), swim bladder area (SBA), eye length (EL), 179 eye width (EW), head width (HW), inter-ocular distance (IOD) and eye-snout distance 180 (ESD) (Figure 1C). For immunochemical determinations, fixed eleutheroembryos were 181 washed several times with PBS, gradually transferred to methanol 100 %, and stored at 182 -80 °C for at least 24h. After rehydration of the larvae, a 50 min depigmentation step (3% H<sub>2</sub>O<sub>2</sub> and 1% KOH in water) was included before the permeabilization steps. 183 184 Whole mount immunohistochemistry was performed as previously described with 185 minor modifications (Thienpont et al., 2011). Larvae were then incubated overnight at 186 4 °C in either anti-parvalbumin monoclonal antibody (Sigma-Aldrich, St. Louis, MO) at 187 1:2000 and anti-acetylated alpha-tubulin monoclonal antibody (Sigma-Aldrich, St. 188 Louis, MO) at 1:1000. The above primary antibodies were used to detect ciliated and 189 microvillous olfactory sensory neurons (anti-parvalbumin, Parv) and cilia from 190 olfactory epithelium and neuromast hair cells (acetylated alpha-tubulin,  $\alpha$ -AT), 191 respectively. The secondary antibodies used were Alexa Fluor 488 and 555 goat anti-192 rabbit IgG and goat anti-mouse IgG (1:300; Molecular Probes).

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#### **2.4.** RNA extraction, library construction and high-throughput sequencing

195 Total RNA was isolated from independent pools of 10 eleutheroembryos (three 196 pools for each condition, separately exposed and treated) using AllPrep DNA/RNA Mini 197 Kit (Qiagen, CA, USA) as described by the manufacturer. Extracted RNA was 198 reconstituted in RNAse-free water and its quantity and quality were determined by a 199 Qubit® RNA BR Assay kit (Thermo Fisher Scientific) and RNA 6000 Nano Assay on a 200 Bioanalyzer 2100 (Agilent Technologies), respectively. RNA preparations were sent to 201 the National Center for Genomic Analysis (CNAG, Barcelona, Spain) for high-202 throughput sequencing (RNA-Seq). All replicates showed RNA concentrations between 203 50 and 200 ng/ $\mu$ L, were free of genomic DNA and had an RNA integrity number (RIN) > 204 8. The RNA-Seq libraries were prepared using KAPA Stranded mRNA-Seq Kit Illumina® 205 Platforms (Kapa Biosystems) with minor modifications. A poly-A based mRNA

206 enrichment with oligo-dT magnetic beads was performed over 500 ng of total RNA as 207 input material. The mRNA was fragmented (resulting RNA fragment size: 80-250nt; 208 major peak: 130nt). The second strand cDNA synthesis was carried out in the presence 209 of dUTP instead of dTTP, to enhance strand specificity. The blunt-ended double-210 stranded cDNA was 3' adenylated and Illumina indexed adapters (Illumina) were 211 ligated. Ligation product was enriched with 15 PCR cycles and validated on an Agilent 212 2100 Bioanalyzer with the DNA 7500 kit. Each final library was sequenced using TruSeq 213 SBS Kit v3-HS (paired-end mode; 2x76bp as read length). An average of 39 million 214 paired-end reads for each sample was generated in a fraction of a sequencing lane on 215 HiSeq 2000 (Illumina). Image analysis, base calling and quality scoring of the run were 216 processed using the manufacturer's software Real Time Analysis (RTA 1.13.48) and 217 FASTQ sequence files were generated by the sequencing analysis software CASAVA. 218 Obtained reads mapped properly to the reference genome in more than a 95%. The 219 majority mapped to exonic regions and to protein-coding genes, with a total of 24,425 220 genes detected. The transcriptomic data discussed in this publication have been 221 deposited in NCBI's Gene Expression Omnibus (Edgar et al., 2002) and are accessible 222 through GEO Series number accession GSE125072 (https://www.ncbi.nlm.nih.gov/geo/query/acc.cgi?acc=GSE125072). A full description 223 224 of mapping quality statistics can be found in **Supplementary Table ST1**.

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#### 226 2.5. Data analysis

#### 227 **2.5.1.** Morphometric statistical analysis

228 Differences between experimental groups in survival, hatching and swim bladder 229 inflation rates (sections 2.2.2. and 2.2.3.) and morphological measurements (section 230 2.3.) were analyzed by non-parametric Kruskal-Wallis tests plus Dunn's pairwise 231 multiple comparisons (significance level at p < 0.05, Bonferroni correction). SPSS 24.0 232 (Armonk, NY: IBM Corp., 2016) was used to carry out statistical tests whereas 233 GraphPad Prism (v. 6.07, GraphPad Software, La Jolla, CA, USA) was used to perform 234 the graphs (Figure 1C and Supplementary Figures SF1, SF2, SF3).

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#### 236 2.5.2. RNA-Seq data analysis

237 RNA-Seq reads were aligned to the D. rerio reference genome (GRCz10) using 238 the STAR software version 2.5.1b (Dobin et al., 2013). The quantification of the genes 239 annotated in GRCz10.84 were performed using RSEM version 1.2.28 (Li and Dewey, 240 2011) with default parameters. Data normalization was carried out using the DESeq2 241 (v.1.10.1) R package (Li and Dewey, 2011; Love et al., 2014), which uses a variant of 242 scaling factor normalization based on the assumption that most genes are not 243 differentially expressed. Differential expression analysis between all experimental 244 conditions was analyzed using the ANOVA-PLS (Analysis of Variance-Partial Least 245 Square) analysis using the Imdme package in R v. 1.0.136, R Core Team (Fresno et al., 246 2014). First, ANOVA decomposed the transcriptomic data matrix of all samples 247 (normalized and scaled) through a linear model that considers the experimental design 248 (PFOS treatment). Secondly, a PLS regression model was built between the matrices 249 obtained in this linear decomposition (X) and a vector (y) that defined the class 250 membership of the samples (control, 0.03, 0.3 and 1.0 mg/L). The analysis identified 251 the features (variables) that described best the differences between groups and it 252 determined if the experimental groups were different from each other. ANOVA-PLS 253 was performed on the normalized data scaled to the control set and log2 transformed, 254 considering each one of the PFOS concentrations (including controls) as a class. Genes 255 showing significant variations among the classes ( $p \le 0.05$ ; 1434 transcripts in total) 256 were selected as DEGs (differentially expressed genes) for further analysis. Hierarchical 257 and PAM (partition around medoids) clustering analysis were performed using the 258 packages gplots, fpc, and cluster in R. The PAM implementation in these packages 259 allowed performing a principal components analysis (PCA) to analyze the covariance 260 matrix of the entered variables and to produce a 2D plot showing the goodness of 261 separation of the defined clusters

262 (https://www.rdocumentation.org/packages/cluster/versions/2.0.6/topics/pam).

Statistically significant differences between genes included in each cluster were assessed by one-way ANOVA followed by post hoc Tukey's B tests (p < 0.05) using the *foreign* and *agricolae* packages in R. Further graphs were carried out using the *gplots* package, also in R environment. DAVID Bioinformatic Resources 6.8 was used for the functional analysis of DEGs. Gene enrichment analysis was estimated in DAVID using the default zebrafish background setting the enrichment significances to a false discovery ratio (FDR) ≤ 5%. Identified modules with at least four hits were included in
the network analysis, using the *reshape2* and *igraph* packages in R
(R\_Development\_Core\_Team, 2008). Graphs were elaborated from an incidence table
of genes (represented by their official gene names, ZFIN.org) using the *igraph* package.
Any two given genes were considered linked if they shared at least one common KEGG
or GO (Gene Ontology) module. Metabolic pathways were obtained from the KEGG
(Kyoto Encyclopedia of Genes and Genomes) database.

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#### 277 **2.5.3.** Benchmark dose and point of departure determination

278 Benchmark doses (BMD) for both morphometric and RNAseq data were 279 calculated using the **BMDExpress** 2.2 sofware 280 (https://www.sciome.com/bmdexpress/)(Phillips et al., 2018; Yang et al., 2007). 281 Benchmark dose lower confidence limits (BMDLs) were calculated to stablish a 282 reference dose or point of departure (PoD) for both datasets (Bhat et al., 2013; EPA, 283 2012; Farmahin et al., 2017; Webster et al., 2015).

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#### 285 3. Results and Discussion

#### 286 **3.1** Survival and morphometric analyses

287 No statistically significant differences (non-parametric Kruskal-Wallis test, p > 288 0.05) were observed in hatching rates at PFOS concentrations up to 100 mg/L 289 (Supplementary Figure SF1A). On the other hand, we observed a significant increase in 290 mortality rates at 5 dpf when embryos were exposed to 7.5, 10 and 100 mg/L and 291 estimated LC50 at 9.1 mg/L. Swim bladder inflation was reduced at PFOS 292 concentrations equal or higher than 2.5 mg/L, at both 4 and 5 dpf. Other lethal 293 endpoints such as lack of heartbeat, coagulated embryos, lack of somite formation, or 294 non-detachment of the tail, were not observed.

Several sublethal morphological changes, like appearance of spinal deformities (bent spine) and scoliosis, reduction of body length (BL), reduction of eye-snout distance (ESD), decrease in swim bladder area (SBA) or increases in yolk sac area (YSA) were observed at 5 dpf at PFOs concentrations equal or higher than 5 mg/L (Figures 1 A-C, **Supplementary Table ST2**). This result was in concordance with previous reports of skeletal deformations (scoliosis and kyphosis) induced by PFOS and other

301 perfluorinated compounds in zebrafish embryos (Hagenaars et al., 2014). Other 302 quantitative morphological parameters, like eye and head width (EW and HW, 303 respectively) and inter-ocular distance (IOD), were reduced respect to the controls at 304 concentrations equal or higher than 2.5 mg/L of PFOS (6-15%, 9-20% and 15-35% 305 respectively, Figures 1 B, C), whereas reductions of the head-trunk angle (HTA) and eye 306 length (EL) showed some significant variations at the 1 mg/L of PFOS concentration 307 (Figures 1 B, C). Note that morphological parameters could not be determined at the 308 highest concentrations (10 and 100 mg/L) due to their high rates of mortality (82 and 309 100 %, respectively). Using the BMDExpress software we calculated PFOS BMDLs for all 310 measured morphological features and found HTA and YSA as the most and less affected parameters, respectively (BMDL<sub>HTA</sub> = 1.27 mg/L, BMDL<sub>YSA</sub> = 5.78 mg/L). Taking 311 312 this into account, we used the median BMDL of all morphological features to estimate 313 2.53 mg/L of PFOS as the morphologic point of departure (PoD) for our study. These 314 results were consistent with previously reported LOAEC values of 1.0 mg/L for 315 activity/behavior (Spulber et al., 2014), 1.0-2.0 mg/L for morphometric alterations 316 (Hagenaars et al., 2014; Shi et al., 2008) or 1.66 mg/L for PFOS ototoxicity (Stengel et 317 al., 2017). We also tested effects on olfactory epithelium and lateral line neuromasts, 318 since they are considered as sensitive markers for neurotoxic or cytotoxic effects of 319 PFOS and other toxicants in zebrafish (Chen and Reese, 2013; de Esch et al., 2012; EPA, 320 2018; Hirose et al., 2011; Sato et al., 2009). Using immunochemical techniques, we 321 observed no evident effects on the olfactory epithelium or on lateral line neuromasts 322 at PFOS concentrations up to 1.0 mg/L, indicating lack of neurotoxicity at this PFOS 323 concentration range (Figure 1D). Taking into consideration all the above exposed 324 reasons, we selected 1.0 mg/L as the maximum PFOS concentration used for the 325 exposures in the transcriptomic study.

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#### 327 3.2 Transcriptome analyses

The ANOVA-PLS analysis identified 1434 transcripts as DEGs in at least one of the exposure groups with respect to the others. Hierarchical clustering of DEGs showed that the expression profiles of the selected DEGs closely reflected the experimental setup, as the different biological replicates for all treatment groups fell into the same hierarchical cluster (Figure 2A, note that the white cells corresponds to

333 the averaged control levels and red and blue cells, correspond to over- and under-334 expressed transcripts, respectively). Unexposed control samples appeared as a clearly 335 differentiated group from the exposed ones. PAM clustering confirmed this rather 336 gradual dose-response pattern by defining two clusters, A and B, corresponding to 337 genes whose abundances decreased (767 genes) or increased (667 genes) upon PFOS 338 exposure, respectively (Figure 2 B-D). Note that the samples of each exposure group 339 clustered together in both clusters. Cluster A showed a clear separation between 340 control and exposed samples, whereas the two highest doses (0.30 and 1.0 mg/L) 341 separated from the 0.03 mg/L group. Cluster B shows a slightly different grouping of 342 samples, as controls and the low dose (0.03 mg/L) clustered separately from the 0.30 343 and 1.0 mg/L groups. Figure 2D shows the distribution of normalized expression values 344 (mean = 0, standard deviation = 1) for all genes in both clusters and reflects the higher 345 difference in the under-expression of the genes (cluster A) between non-treated 346 (control) and treated samples than those among the different dose groups. The dose-347 response pattern is less evident in cluster B. We thus conclude that the LOAEC for 348 transcriptomic effects of PFOS (at least, for the major part of it) is likely below 0.03 349 mg/L (30 ppb), as these samples separated themselves from controls. Moreover, we 350 estimated 0.011 mg/L of PFOS (11 ppb) as PFOS transcriptomic PoD, about a third of 351 the lowest concentration used in the transcriptomic study (Supplementary Figure 352 SF1B). This is more than 200 fold lower than the estimated PoD value for 353 morphometric changes (Supplementary Figure SF1B), and lower than the reported 354 limit of detection for transcriptomic effects of PFOS by targeted analyses (0.10 mg/L, 355 (Jantzen et al., 2016; Shi et al., 2008). RNA-seq transcriptomic data exposed in this 356 study was confirmed by RT-qPCR (Supplementary Methods, Supplementary Figure 357 SF2, Supplementary Table ST3) showing a strong correlation between both techniques in several selected genes ( $r^2 = 0.814$ , p <  $10^{-4}$ ). 358

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#### 360 **3.3 Functional analysis**

361 DAVID functional analysis showed a relatively small subset of functional classes 362 significantly enriched either in any of the two clusters or in the whole DEGs subset 363 (Table 1, see also Figure 3A and supplementary Table ST4). Particularly notable is the 364 case for over-represented transcripts (Cluster B), in which all significantly enriched

365 functional classes were related to lipid transport and/or metabolism (Table 1). This is 366 consistent with the known disruption of lipid homeostasis by PFOS and other 367 structurally-related compounds in vertebrates (Das et al., 2017; Huang et al., 2016), 368 and it shows a similar pattern as the lipid metabolism/transport transcriptomic effects 369 induced by other EDCs like Bisphenol A (Martínez et al., 2018) (Supplementary Figure SF3,  $r^2 = 0.678$ ,  $p < 10^{-4}$ ). The presence of yolk sac remains observed in the present 370 371 study (section 3.2.) is also in agreement with the alterations in the lipid homeostasis 372 shown at the transcriptomic level.

373 Several functional modules associated to Cluster A were identified by DAVID 374 analysis, although most of them corresponded to structural, rather than functional 375 categories (Table 1, Figure 3A). For example, the observed transcriptomic 376 dysregulations in myosin, actin and tropomyosin transcripts (Supplementary figure SF4) are in agreement with the observed spinal deformities (section 3.2.), which have 377 378 been in turn related to alterations in myosin and other muscle fibers (Huang et al., 379 2010). As it can be observed in **Supplementary figure SF4**, genes in cluster A exhibited 380 a classic dose-response pattern, increasing their expression according to the PFOS 381 exposure. Nevertheless, the relationship between the dysregulation of these genes 382 and the effects on the spinal cord and tail structure need to be confirmed by a further 383 specific research.

384 Most enriched functional categories in Cluster A (KEGG or GO\_BP terms) 385 appeared related to the immunological system and/or pathogen response, including 386 several cytokines, cytokine signaling-related proteins, proteases implicated to antigen 387 presentation, or the JAK-STAT signaling pathway (Table 1). This is consistent with the 388 known dysregulation of the immune system caused by PFOS and other perfluorinated 389 substances, which includes changes in cytokines production (DeWitt et al., 2009; Lau et 390 al., 2007; Peden-Adams et al., 2008; Suo et al., 2017). Structurally-defined modules (INTERPRO or GO\_MF terms in Table 1) included a relatively large group of genes 391 392 associated to the structural domain B30.2/SPRY and Zn-ion binding proteins, among 393 others. While the structural domain B30.2/SPRY have been reported to be involved in 394 immunity and immunoglobulins (D'Cruz et al., 2013; Howe et al., 2016; Woo et al., 395 2006), zinc fingers (structural motif of the proteins which Zn binds to) have an 396 extraordinary diversity of both structures and functions (Kluska et al., 2018; Laity et al.,

397 2001).

398 Considering the whole DEG dataset, more general functions appear, including 399 signaling, extracellular space and transmembrane receptor activity. Inspection of the 400 genes identified by their structural, rather than functional characteristics, revealed the 401 biological processes potentially affected by their de-regulation. These 402 interdependences can be visualized in the network shown in Figure 3B. For example, 403 the structural class "IPR001507: Zona pellucida domain" shares some genes with the 404 functional class "cell-matrix adhesion" (both in cluster A), suggesting alterations in cell-405 cell interactions. Similarly, the structural class "IPR011029: Death-like domain" may be 406 related to programmed cell-death pathways, as most of the 13 genes identified as 407 DEGs are also related to caspase recruitment, apoptosis, or TNF (tumor necrosis factor) 408 activation (Table 1). This is consistent with previous studies in cells (Cui et al., 2017; Shi 409 et al., 2008; Zhang et al., 2013) and zebrafish (Shi et al., 2008). Particularly noticeable is 410 the structural-based module "IPR001870:B30.2/SPRY domain" that, together with 411 closely-related structural terms (IPR003879 and IPR013320, among others), includes 412 many components of the so-called TRIM protein family, a group of E3-ligase proteins 413 functionally related to the innate immunity development (Versteeg et al., 2013) (Table 414 1). Most detected members of this family corresponded to the fish-specific finTRIM 415 subfamily (FTR genes, Table 1), also putatively related to different aspects of the fish 416 immune response (van der Aa et al., 2012). Also members of the TRIM family are the 417 bloodthirsty (btr) genes, orthologous to the human TRIM39 protein and also 418 potentially related with the above mentioned immune response (Luo et al., 2017; van 419 der Aa et al., 2012; Zhang et al., 2015). All these genes appeared as down-regulated by 420 the exposure to PFOS (Figure 3, note that they are placed in cluster A).

421 Cell adhesion, another of the affected pathways in our study, is mediated by 422 four major protein superfamilies (immunoglobulins, selectins, cadherins and integrins; 423 a.k.a. CAMs). Immunoglobulins and selectins are involved in the immune response 424 (Mashoof and Criscitiello, 2016; Roca et al., 2008; Sun et al., 2015); cadherins in the 425 actin fibers linkeage through catenins (Li-Villarreal et al., 2016); and integrins facilitate 426 the extracellular matrix adhesion, also activating the signaling transduction pathways 427 including the apoptosis signals, among others (Mould et al., 2006). We observed that 428 all those pathways were transcriptionally affected by PFOS (Table1, Figure 3). Although the data suggests the existence of a PFOS-induced initial toxic event related to cell
adhesion and/or the signaling pathways, further studies are required in order to
achieve a deeper comprehension about its mode of action.

432 It is remarkable how interweaved were the observed transcriptomics effects 433 with themselves and with the phenotypic alterations, suggesting a common initial 434 triggering point. For example, the structural and functional diversity of lectins (the 435 above mentioned IPR013320 structural term is a subgroup of them) have been 436 previously studied in teleost fish, showing their involvement not only in the immune 437 response (Vasta et al., 2011), but also in lipid regulation (Cambi et al., 2005; Ng et al., 438 1989), and that lectin malfunction may result in bent tails and other skeletal muscle 439 problems in zebrafish (Ahmed et al., 2009), which are similar to the ones observed in 440 our exposed animals.

441

#### 442 3.4 Toxicological relevance

443

444 Many animal and in vitro human systems indicate that exposure to PFOS may result in 445 neurotoxicity, immunotoxicity, thyroid disruption, reproductive, cardiovascular and 446 pulmonary toxicity, and diverse toxic effects in liver and kidneys (Das et al., 2017; 447 DeWitt et al., 2009; Lau et al., 2007; Suo et al., 2017; Zeng et al., 2019). These studies 448 suggest that PFOS represent a significant hazard for human health. This has been at 449 least partially confirmed by the still scarce epidemiological studies, being water and 450 food the two main sources of exposure (Zeng et al., 2019). In addition to their 451 implication for human health, the wide use, high persistence, and bioaccumulative 452 properties of PFOS and other PFAS implicate that they are ubiquitous in the 453 environment, and in particular, in aquatic bodies. PFOS and other PFAS has been 454 detected in fish and marine mammal and bird tissues at ng-µg/g ww levels (Ahrens and 455 Bundschuh, 2014; Giesy and Kannan, 2001; Houde et al., 2011; Lau et al., 2007; 456 Rodriguez-Jorquera et al., 2016). Both molecular and macroscopic effects have been 457 reported for PFOS and other perfluoroorganic compounds in zebrafish embryos and 458 adults (Chen et al., 2013; Cui et al., 2017; Fai Tse et al., 2016; Hagenaars et al., 2014; 459 Lau et al., 2007; Shi et al., 2009). Similar effects have been reported in fathead 460 minnows (*Pimephales promelas*) exposed to urban wastewaters presenting from ng to

μg/L levels of different PFAS (Rodriguez-Jorquera et al., 2015). In addition, there is a
growing evidence of toxic effects of perflouroorganic compounds in invertebrates and
plants (Giesy et al., 2010; Li, 2009; Stylianou et al., 2019).Therefore, we consider that
our findings reflect toxic effects relevant not only for human health, but also for the
environment.

466 Lipid metabolism, immunological response and transmembrane/intracellular 467 signaling appeared as the main cellular functions affected by PFOS exposure at the 468 molecular level (Das et al., 2017; DeWitt et al., 2009; Lau et al., 2007; Suo et al., 2017). 469 The favorite mechanism of action (MoA) proposed for PFOS toxicity involves its 470 interaction with nuclear receptors intimately involved in metabolic regulation and 471 immunological functions, like PPAR $\alpha$  (peroxisome proliferator activating receptor), 472 CAR (constitutive and rostane receptor), PXR (pregnane X receptor) or FXR (farnesoid X 473 receptor) (Lau et al., 2007; White et al., 2011). While our experimental approach does 474 not allow the characterization of the molecular mechanisms underlying the observed 475 changes, they are essentially consistent with this MoA, which explain the simultaneous 476 deregulation of genes involved in lipid metabolism and immuno response, among 477 others. The effect of PFOS on lipid metabolism (including glycerophospholipids) of 478 zebrafish embryos was also observed at the metabolic level (Ortiz-Villanueva et al., 479 2018), and it is likely related to the observed changes in yolk sac absorption. Therefore, 480 our data show essentially the same kind of responses at gene expression, metabolism 481 and morphological level.

482 PFOS is usually found at low levels (in the range of ng/L) in surface waters 483 (Kunacheva et al., 2011; Vedagiri et al., 2018), particularly those under the influence of 484 sewage treatment plants (Rodriguez-Jorquera et al., 2015; Rodriguez-Jorquera et al., 485 2016). However, PFOS is considered a persistent pollutant and its high 486 bioaccumulation and low elimination rates (Huang et al., 2010; Kannan et al., 2005) 487 cause PFOS to be found at very high levels (0.05-5.0 mg/kg) in wildlife (Hoff et al., 488 2005). It has been reported a 10-fold bioaccumulation between the water and 489 exposed-whole zebrafish larvae tissues in only 5 days of PFOS exposure (Huang et al., 490 2010). Our PoD values for PFOS estimated from either morphometric (2.53 mg/L) or transcriptional data (0.011 mg/L) were higher than the environmental limits 491 established for drinking water  $(10^{-5}-10^{-4} \text{ mg/L})$  according to both EPA and EFSA, as well 492

493 as other regulatory agencies (EC, 2004; Grandjean, 2018). Nevertheless, exposure 494 routes for PFOS are diverse, and there are considerably higher legal limits for some 495 purposes, like the 10 mg/kg limit for substances and preparations in the EU (EC, 2004). 496 Regarding internal exposure levels, several studies have observed concentrations from 497 0.002 mg/L to 0.080 mg/L of PFOS in human serum of individuals from Asia, North 498 America and Europe (Alexander et al., 2008; Jin et al., 2007; Vedagiri et al., 2018; Zeng 499 et al., 2015) and even higher concentrations (0.145-0.381 mg/kg) in fish eggs (Kannan 500 et al., 2005). This indicates that both humans and wildlife may be exposed to PFOS 501 concentrations similar to or even higher than our calculated transcriptomic PoD value. 502 Barring the need for dose-conversion and pharmacokinetics modeling to extrapolate 503 the effects in zebrafish to humans, our results supported a very low safety margin, as 504 they could imply that the immune system of exposed individuals, among others systems, could be affected at least at the transcriptomic level. In this regard, the 505 506 suppression of immunological responses by different pollutants has been largely 507 neglected until very recently, partially because the lack of an appropriate animal or cell 508 models for its toxicological assessment (Möller et al., 2014; Rehberger et al., 2017; 509 Segner et al., 2017). However, there is an increasing amount of evidence that some 510 pollutants may indeed decrease the ability of fish and other vertebrates to fend off 511 challenging infections (Fang et al., 2013; Quesada-Garcia et al., 2016). In the same 512 direction, a report from the USA National Toxicology program also identified 513 immunotoxicity as an emerging adverse effect of perfluorinated substances in humans 514 (NPT, 2016). For all those reasons, we propose the inclusion of immunotoxicity tests in 515 the existing zebrafish embryo testing schemes, which currently cover cardiovascular, 516 nervous, neuromuscular, gastrointestinal and thyroid systems (Raldúa and Piña, 2014), 517 to assess toxic effects of different substances in humans and other vertebrates.

518

#### 519 4. Conclusions

520 Our results suggest a complex, multiple endocrine disruption-like toxic effects at 521 concentrations well below the LOAEC/NOAEC for many of the macroscopic effects 522 traditionally linked to PFOS toxicity in zebrafish embryos. While our results confirm the 523 known effect of PFOS in the spinal cord, and its potential role as lipid disruptor, we 524 found a significant decrease in the expression of many genes related to natural

immunity and defense against infections, previously reported in other organisms. 525 526 Therefore, we suggest that a common initial key event may trigger all observed 527 adverse effects elicited by exposure to PFOS. We propose that the transcriptional 528 pattern may be a marker for the immunotoxic effects of PFOS and other related 529 substances in fish and other vertebrates, including humans. As the estimated PoD 530 values for transcriptional changes occurred at concentrations already found in living 531 organisms, including humans, our data suggest that current maximal tolerable levels 532 may not protect adequately environmental and human health and might need to be 533 revised.

534

#### 535 Conflicts of interest

536 The authors declare that they have no conflicts of interest.

537

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#### 552 References

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- Ahmed, H., Du, S.J., Vasta, G.R., 2009. Knockdown of a galectin-1-like protein in
- zebrafish (Danio rerio) causes defects in skeletal muscle development, 3 ed, pp. 277-283.
- Ahrens, L., Bundschuh, M., 2014. Fate and effects of poly- and perfluoroalkyl
- substances in the aquatic environment: A review. Environmental Toxicology andChemistry 33, 1921-1929.
- 561 Alexander, J., Atli Auðunsson, G., Benford, D., Cockburn, A., Cravedi, J.-P., Dogliotti, E.,
- 562 Di Domenico, A., Luisa Fernández-Cruz, M., Fink-Gremmels, J., Fürst, P., Galli, C.,
- 563 Grandjean, P., Gzyl, J., Heinemeyer, G., Johansson, N., Mutti, A., Schlatter, J., van
- Leeuwen, R., van Peteghem, C., Verger, P., 2008. Perfluorooctane sulfonate (PFOS),
- 565 perfluorooctanoic acid (PFOA) and their salts Scientific Opinion of the Panel on 566 Contaminants in the Food chain. Efsa Journal 6, 1-131.
- 567 Ankley, G.T., Kuehl, D.W., Kahl, M.D., Jensen, K.M., Linnum, A., Leino, R.L., Villeneuve,
- 568 D.A., 2005. Reproductive and developmental toxicity and bioconcentration of
- 569 perfluorooctanesulfonate in a partial life-cycle test with the fathead minnow
- 570 (Pimephales promelas). Environmental Toxicology and Chemistry 24, 2316-2324.
- 571 Arvaniti, O.S., Stasinakis, A.S., 2015. Review on the occurrence, fate and removal of
- 572 perfluorinated compounds during wastewater treatment. Science of the Total573 Environment 524, 81-92.
- 574 Avdesh, A., Chen, M., Martin-Iverson, M.T., Mondal, A., Ong, D., Rainey-Smith, S.,
- Taddei, K., Lardelli, M., Groth, D.M., Verdile, G., Martins, R.N., 2012. Regular Care and
  Maintenance of a Zebrafish (<em>Danio rerio</em>) Laboratory: An Introduction.
  Journal of Visualized Experiments, e4196-e4196.
- 578 Baker, M.E., Hardiman, G., 2014. Transcriptional analysis of endocrine disruption using
- zebrafish and massively parallel sequencing. Journal of Molecular Endocrinology 52,R241-256.
- 581 Bhat, V.S., Hester, S.D., Nesnow, S., Eastmond, D.A., 2013. Concordance of
- Transcriptional and Apical Benchmark Dose Levels for Conazole-Induced Liver Effects inMice. Toxicological Sciences 136, 205-215.
- 584 Caballero-Gallardo, K., Olivero-Verbel, J., Freeman, J.L., 2016. Toxicogenomics to
- evaluate endocrine disrupting effects of environmental chemicals using the zebrafishmodel. Current genomics 17, 515-527.
- 587 Cambi, A., Koopman, M., Figdor, C.G., 2005. How C-type lectins detect pathogens, pp.588 481-488.
- 589 Chaparro-Ortega, A., Betancourt, M., Rosas, P., Vázquez-Cuevas, F.G., Chavira, R.,
- 590 Bonilla, E., Casas, E., Ducolomb, Y., 2018. Endocrine disruptor effect of perfluorooctane
- sulfonic acid (PFOS) and perfluorooctanoic acid (PFOA) on porcine ovarian cell
- 592 steroidogenesis. Toxicology in Vitro 46, 86-93.
- 593 Chen, J., Das, S.R., La Du, J., Corvi, M.M., Bai, C., Chen, Y., Liu, X., Zhu, G., Tanguay, R.L.,
- 594 Dong, Q., Huang, C., 2013. Chronic PFOS exposures induce life stage-specific behavioral
- 595 deficits in adult zebrafish and produce malformation and behavioral deficits in F1
- 596 offspring. Environmental Toxicology and Chemistry 32, 201-206.
- 597 Chen, J., Tanguay, R.L., Tal, T.L., Gai, Z., Ma, X., Bai, C., Tilton, S.C., Jin, D., Yang, D.,
- 598 Huang, C., Dong, Q., 2014. Early life perfluorooctanesulphonic acid (PFOS) exposure
- 599 impairs zebrafish organogenesis. Aquatic Toxicology 150, 124-132.

- 600 Chen, Y., Reese, D.H., 2013. A Screen for Disruptors of the Retinol (Vitamin A) Signaling
- Pathway. Birth Defects Research Part B-Developmental and Reproductive Toxicology98, 276-282.
- 603 Corsini, E., Avogadro, A., Galbiati, V., dell'Agli, M., Marinovich, M., Galli, C.L., Germolec,
- D.R., 2011. In vitro evaluation of the immunotoxic potential of perfluorinated
- 605 compounds (PFCs). Toxicology and Applied Pharmacology 250, 108-116.
- 606 Cui, Y., Lv, S., Liu, J., Nie, S., Chen, J., Dong, Q., Huang, C., Yang, D., 2017. Chronic
- 607 perfluorooctanesulfonic acid exposure disrupts lipid metabolism in zebrafish. Human &
  608 Experimental Toxicology 36, 207-217.
- 609 D'Cruz, A.A., Babon, J.J., Norton, R.S., Nicola, N.A., Nicholson, S.E., 2013. Structure and
- 610 function of the SPRY/B30.2 domain proteins involved in innate immunity. Wiley-611 Blackwell, pp. 1-10.
- Das, K.P., Wood, C.R., Lin, M.T., Starkov, A.A., Lau, C., Wallace, K.B., Corton, J.C.,
- 613 Abbott, B.D., 2017. Perfluoroalkyl acids-induced liver steatosis: Effects on genes
- 614 controlling lipid homeostasis. Toxicology 378, 37-52.
- de Esch, C., Slieker, R., Wolterbeek, A., Woutersen, R., de Groot, D., 2012. Zebrafish as
- 616 potential model for developmental neurotoxicity testing: a mini review.
- 617 Neurotoxicology and Teratology 34, 545-553.
- 618 DeWitt, J.C., Shnyra, A., Badr, M.Z., Loveless, S.E., Hoban, D., Frame, S.R., Cunard, R.,
- Anderson, S.E., Meade, B.J., Peden-Adams, M.M., Luebke, R.W., Luster, M.I., 2009.
- 620 Immunotoxicity of perfluorooctanoic acid and perfluorooctane sulfonate and the role621 of peroxisome proliferator-activated receptor alpha, pp. 76-94.
- 622 DeWitt, J.C., Williams, W.C., Creech, N.J., Luebke, R.W., 2016. Suppression of antigen-
- 623 specific antibody responses in mice exposed to perfluorooctanoic acid: Role of PPAR
- alpha and T- and B-cell targeting. Journal of Immunotoxicology 13, 38-45.
- 625 Dobin, A., Davis, C.A., Schlesinger, F., Drenkow, J., Zaleski, C., Jha, S., Batut, P.,
- 626 Chaisson, M., Gingeras, T.R., 2013. STAR: Ultrafast universal RNA-seq aligner.627 Bioinformatics 29, 15-21.
- 628 Du, G., Hu, J., Huang, H., Qin, Y., Han, X., Wu, D., Song, L., Xia, Y., Wang, X., 2013.
- Perfluorooctane sulfonate (PFOS) affects hormone receptor activity, steroidogenesis,and expression of endocrine-related genes in vitro and in vivo. Environmental
- 631 Toxicology and Chemistry 32, 353-360.
- 632 EC, 2004. Regulation (EC) No 850/2004 of the European Parliament and of the Council
- 633 of 29 April 2004 on persistent organic pollutants and amending Directive 79/117/EEC,
- 634 in: Council, E.P.a. (Ed.). European Commission, Brussels (Belgium).
- 635 Edgar, R., Domrachev, M., Lash, A.E., 2002. Gene Expression Omnibus: NCBI gene
- expression and hybridization array data repository. Nucleic Acids Res 30, 207-210.
- 637 EPA, R.A.F., 2012. Benchmark Dose Technical Guidance, in: Agency, U.E.P. (Ed.),
- 638 EPA/100/R-12/001. EPA, Washington, D.C.
- 639 EPA, R.A.F., 2018. Exposure to PFOS, PFHxS, or PFHxA, but not GenX, Nafion BP1, or
- ADONA, elicits developmental neurotoxicity in larval zebrafish, in: Agency, U.S.E.P.(Ed.).
- 642 European Commission, D.E., 2000. Towards the establishment of a priority list of
- 643 substances for further evaluation of their role in endocrine disruption.
- Fai Tse, W.K., Li, J.W., Kwan Tse, A.C., Chan, T.F., Hin Ho, J.C., Sun Wu, R.S., Chu Wong,
- 645 C.K., Lai, K.P., 2016. Fatty liver disease induced by perfluorooctane sulfonate: Novel
- 646 insight from transcriptome analysis. Chemosphere 159, 166-177.

- Fang, C., Huang, Q.S., Ye, T., Chen, Y.J., Liu, L.P., Kang, M., Lin, Y., Shen, H.Q., Dong, S.J.,
- 648 2013. Embryonic exposure to PFOS induces immunosuppression in the fish larvae of
  649 marine medaka. Ecotoxicology and Environmental Safety 92, 104-111.
- 650 Farmahin, R., Williams, A., Kuo, B., Chepelev, N.L., Thomas, R.S., Barton-Maclaren, T.S.,
- 651 Curran, I.H., Nong, A., Wade, M.G., Yauk, C.L., 2017. Recommended approaches in the
- application of toxicogenomics to derive points of departure for chemical risk
- assessment. Archives of Toxicology 91, 2045-2065.
- Fresno, C., Balzarini, M.G., Fernandez, E.A., 2014. Imdme: Linear Models on Designed
  Multivariate Experiments in R. Journal of Statistical Software 56, 1-16.
- 656 Giesy, J.P., Kannan, K., 2001. Global distribution of perfluorooctane sulfonate in
- 657 wildlife. Environmental Science & Technology 35, 1339-1342.
- Giesy, J.P., Naile, J.E., Khim, J.S., Jones, P.D., Newsted, J.L., 2010. Aquatic Toxicology of
- 659 Perfluorinated Chemicals, in: Whitacre, D.M. (Ed.), Reviews of Environmental
- 660 Contamination and Toxicology, Vol 202. Springer, New York, pp. 1-52.
- 661 Grandjean, P., 2018. Delayed discovery, dissemination, and decisions on intervention
- in environmental health: a case study on immunotoxicity of perfluorinated alkylatesubstances. Environmental Health 17, 62-62.
- Hagenaars, A., Stinckens, E., Vergauwen, L., Bervoets, L., Knapen, D., 2014. PFOS
  affects posterior swim bladder chamber inflation and swimming performance of
  zebrafish larvae. Aquatic Toxicology 157, 225-235.
- Hill, A.J., Teraoka, H., Heideman, W., Peterson, R.E., 2005. Zebrafish as a model
  vertebrate for investigating chemical toxicity, pp. 6-19.
- Hirose, Y., Simon, J.A., Ou, H.C., 2011. Hair Cell Toxicity in Anti-cancer Drugs: Evaluating
- an Anti-cancer Drug Library for Independent and Synergistic Toxic Effects on Hair Cells
- Using the Zebrafish Lateral Line. Jaro-Journal of the Association for Research in
- 672 Otolaryngology 12, 719-728.
- Hoff, P.T., Van Campenhout, K., Van De Vijver, K., Covaci, A., Bervoets, L., Moens, L.,
- Huyskens, G., Goemans, G., Belpaire, C., Blust, R., De Coen, W., 2005. Perfluorooctane
- 675 sulfonic acid and organohalogen pollutants in liver of three freshwater fish species in
- 676 Flanders (Belgium): relationships with biochemical and organismal effects.
- 677 Environmental Pollution 137, 324-333.
- 678 Houde, M., De Silva, A.O., Muir, D.C.G., Letcher, R.J., 2011. Monitoring of
- 679 Perfluorinated Compounds in Aquatic Biota: An Updated Review PFCs in Aquatic Biota.
  680 Environmental Science & Technology 45, 7962-7973.
- Howe, K., Schiffer, P.H., Zielinski, J., Wiehe, T., Laird, G.K., Marioni, J.C., Soylemez, O.,
- 682 Kondrashov, F., Leptin, M., 2016. Structure and evolutionary history of a large family of
- 683 NLR proteins in the zebrafish. Open Biology 6, 160009-160009.
- Huang, H., Huang, C., Wang, L., Ye, X., Bai, C., Simonich, M.T., Tanguay, R.L., Dong, Q.,
- 2010. Toxicity, uptake kinetics and behavior assessment in zebrafish embryos following
   exposure to perfluorooctanesulphonicacid (PFOS). Aquatic Toxicology 98, 139-147.
- Huang, S.S.Y., Benskin, J.P., Chandramouli, B., Butler, H., Helbing, C.C., Cosgrove, J.R.,
- 688 2016. Xenobiotics Produce Distinct Metabolomic Responses in Zebrafish Larvae (Danio
- rerio). Environmental Science & Technology 50, 6526-6535.
- 690 Jantzen, C.E., Annunziato, K.A., Bugel, S.M., Cooper, K.R., 2016. PFOS, PFNA, and PFOA
- 691 sub-lethal exposure to embryonic zebrafish have different toxicity profiles in terms of
- 692 morphometrics, behavior and gene expression. Aquatic Toxicology 175, 160-170.

- Jensen, A.A., Leffers, H., Emerging endocrine disrupters: Perfluoroalkylated substances,2 ed, pp. 161-169.
- Jin, Y., Saito, N., Harada, K.H., Inoue, K., Koizumi, A., 2007. Historical trends in human
- 696 serum levels of perfluorooctanoate and perfluorooctane sulfonate in Shenyang, China.
  697 The Tohoku journal of experimental medicine 212, 63-70.
- Kannan, K., Tao, L., Sinclair, E., Pastva, S.D., Jude, D.J., Giesy, J.P., 2005. Perfluorinated
- 699 compounds in aquatic organisms at various trophic levels in a Great Lakes food chain.
  700 Archives of Environmental Contamination and Toxicology 48, 559-566.
- Kato, K., Wong, L.-Y., Basden, B.J., Calafat, A.M., 2013. Effect of temperature and
  duration of storage on the stability of polyfluoroalkyl chemicals in human serum.
  Chemosphere 91, 115-117.
- 704 Kato, K., Ye, X., Calafat, A.M., 2015. PFASs in the General Population, pp. 51-76.
- Kimmel, C.B., Ballard, W.W., Kimmel, S.R., Ullmann, B., Schilling, T.F., 1995. Stages of embryonic development of the zebrafish. Dev Dyn 203, 253-310.
- Kluska, K., Adamczyk, J., Krężel, A., 2018. Metal binding properties, stability and
  reactivity of zinc fingers, pp. 18-64.
- 709 Kunacheva, C., Tanaka, S., Fujii, S., Boontanon, S.K., Musirat, C., Wongwattana, T.,
- 710 Shivakoti, B.R., 2011. Mass flows of perfluorinated compounds (PFCs) in central
- 711 wastewater treatment plants of industrial zones in Thailand. Chemosphere 83, 737-712 744.
- 713 Lai, K.P., Ng, A.H.M., Wan, H.L., Wong, A.Y.M., Leung, C.C.T., Li, R., Wong, C.K.C., 2018.
- 714 Dietary Exposure to the Environmental Chemical, PFOS on the Diversity of Gut
- 715 Microbiota, Associated With the Development of Metabolic Syndrome. Frontiers in716 Microbiology 9, 11.
- 717 Laity, J.H., Lee, B.M., Wright, P.E., 2001. Zinc finger proteins: New insights into
- structural and functional diversity. Elsevier Current Trends, pp. 39-46.
- Lau, C., 2015. Perfluorinated compounds: an overview. Humana Press, Cham, pp. 1-21.
- Lau, C., Anitole, K., Hodes, C., Lai, D., Pfahles-Hutchens, A., Seed, J., 2007.
- Perfluoroalkyl acids: A review of monitoring and toxicological findings. ToxicologicalSciences 99, 366-394.
- Li, B., Dewey, C.N., 2011. RSEM: accurate transcript quantification from RNA-Seq data with or without a reference genome. Bmc Bioinformatics 12, 323-323.
- Li, M.H., 2009. Toxicity of Perfluorooctane Sulfonate and Perfluorooctanoic Acid to
- 726 Plants and Aquatic Invertebrates. Environmental Toxicology 24, 95-101.
- 127 Li-Villarreal, N., Forbes, M.M., Loza, A.J., Chen, J., Ma, T., Helde, K., Moens, C.B., Shin,
- 728 J., Sawada, A., Hindes, A.E., Dubrulle, J., Schier, A.F., Longmore, G.D., Marlow, F.L.,
- 729 Solnica-Krezel, L., 2016. Dachsous1b cadherin regulates actin and microtubule
- 730 cytoskeleton during early zebrafish embryogenesis. Development 143, 1832-1832.
- 731 Loos, R., Carvalho, R., António, D.C., Comero, S., Locoro, G., Tavazzi, S., Paracchini, B.,
- 732 Ghiani, M., Lettieri, T., Blaha, L., Jarosova, B., Voorspoels, S., Servaes, K., Haglund, P.,
- 733 Fick, J., Lindberg, R.H., Schwesig, D., Gawlik, B.M., 2013. EU-wide monitoring survey on
- emerging polar organic contaminants in wastewater treatment plant effluents. Water
- 735 Research 47, 6475-6487.
- 736 Love, M.I., Huber, W., Anders, S., 2014. Moderated estimation of fold change and
- dispersion for RNA-seq data with DESeq2. Genome Biology 15, 550-550.

- 738 Luo, K., Li, Y., Xia, L., Hu, W., Gao, W., Guo, L., Tian, G., Qi, Z., Yuan, H., Xu, Q., 2017.
- Analysis of the expression patterns of the novel large multigene TRIM gene family(finTRIM) in zebrafish. Fish & Shellfish Immunology 66, 224-230.
- 741 Lyu, X.-J., Li, W.-W., Lam, P.K.S., Yu, H.-Q., 2015. Insights into perfluorooctane
- sulfonate photodegradation in a catalyst-free aqueous solution. Scientific Reports 5,9353-9353.
- 744 Martínez, R., Esteve-Codina, A., Herrero-Nogareda, L., Ortiz-Villanueva, E., Barata, C.,
- 745 Tauler, R., Raldúa, D., Piña, B., Navarro-Martín, L., 2018. Dose-dependent
- 746 transcriptomic responses of zebrafish eleutheroembryos to Bisphenol A.
- 747 Environmental Pollution.
- 748 Mashoof, S., Criscitiello, M., 2016. Fish immunoglobulins. Biology 5, 45-45.
- 749 Möller, A.-M., Korytář, T., Köllner, B., Schmidt-Posthaus, H., Segner, H., 2014. The
- 750 teleostean liver as an immunological organ: Intrahepatic immune cells (IHICs) in
- 751 healthy and benzo[a]pyrene challenged rainbow trout (Oncorhynchus mykiss).
- 752 Developmental & Comparative Immunology 46, 518-529.
- 753 Mortazavi, A., Williams, B.A., McCue, K., Schaeffer, L., Wold, B., 2008. Mapping and
- 754 quantifying mammalian transcriptomes by RNA-Seq. Nature Methods 5, 621-628.
- 755 Mould, A.P., McLeish, J.A., Huxley-Jones, J., Goonesinghe, A.C., Hurlstone, A.F.L., Boot-
- Handford, R.P., Humphries, M.J., 2006. Identification of multiple integrin  $\beta$ 1 homologs
- in zebrafish (Danio rerio). BMC Cell Biology 7, 24-24.
- 758 Mushtaq, M.Y., Verpoorte, R., Kim, H.K., 2013. Zebrafish as a model for systems
- biology. Biotechnology and Genetic Engineering Reviews, Vol 29, Issue 2 29, 187-205.
- 760 Ng, T.B., Li, W.W., Yeung, H.W., 1989. Effects of lectins with various carbohydrate
- binding specificities on lipid metabolism in isolated rat and hamster adipocytes.
- 762 International Journal of Biochemistry 21, 149-156.
- NPT, National Toxicology Program, 2016. NTP monograph: Immunotoxicity associated
  with exposure to perfluorooctanoic acid or perfluorooctane sulfonate. US Department
  of Human and Health Services.
- 766 OECD, 2002. Co-operation on existing chemicals: hazard assessment of
- 767 perfluorooctane sulfonate (PFOS) and its salts, pp. 1-362.
- 768 OECD, 2013. Test No. 236: Fish Embryo Acute Toxicity (FET) Test. OECD Publishing.
- Olsen, G.W., 2015. PFAS Biomonitoring in Higher Exposed Populations. Toxicological
   Effects of Perfluoroalkyl and Polyfluoroalkyl Substances, 77-125.
- 771 Ortiz-Villanueva, E., Jaumot, J., Martinez, R., Navarro-Martin, L., Pina, B., Tauler, R.,
- 772 2018. Assessment of endocrine disruptors effects on zebrafish (Danio rerio) embryos
- by untargeted LC-HRMS metabolomic analysis. The Science of the total environment635, 156-166.
- Paul, A.G., Jones, K.C., Sweetman, A.J., 2009. A First Global Production, Emission, And
  Environmental Inventory For Perfluorooctane Sulfonate. Environmental Science &
- 777 Technology 43, 386-392.
- 778 Peden-Adams, M.M., Keller, J.M., Eudaly, J.G., Berger, J., Gilkeson, G.S., Keil, D.E.,
- 2008. Suppression of humoral immunity in mice following exposure to perfluorooctanesulfonate. Toxicological Sciences 104, 144-154.
- 781 Phillips, J.R., Svoboda, D.L., Tandon, A., Patel, S., Sedykh, A., Mav, D., Kuo, B., Yauk,
- 782 C.L., Yang, L., Thomas, R.S., Gift, J.S., Davis, J.A., Olysyzk, L., Merrick, B.A., Paules, R.S.,
- 783 Parham, F., Saddler, T., Shah, R.R., Auerbach, S.S., 2018. BMDExpress 2: Enhanced
- 784 transcriptomic dose-response analysis workflow. Bioinformatics (Oxford, England).

- 785 Quesada-Garcia, A., Encinas, P., Valdehita, A., Baumann, L., Segner, H., Coll, J.M.,
- Navas, J.M., 2016. Thyroid active agents T3 and PTU differentially affect immune gene
- transcripts in the head kidney of rainbow trout (Oncorynchus mykiss). AquaticToxicology 174, 159-168.
- 789 R\_Development\_Core\_Team, 2008. R: A language and environment for statistical
- 790 computing. R Foundation for Statistical Computing, Vienna, Austria.
- 791 Raldua, D., Otero, D., Fabra, M., Cerda, J., 2008. Differential localization and regulation
- of two aquaporin-1 homologs in the intestinal epithelia of the marine teleost Sparus
- aurata. American Journal of Physiology-Regulatory Integrative and ComparativePhysiology 294, R993-R1003.
- Raldúa, D., Piña, B., 2014. In vivo zebrafish assays for analyzing drug toxicity. Expert
  Opinion on Drug Metabolism and Toxicology 10, 685-697.
- Rehberger, K., Werner, I., Hitzfeld, B., Segner, H., Baumann, L., 2017. 20 Years of fish
  immunotoxicology what we know and where we are. Critical Reviews in Toxicology
  47, 516-542.
- Reuter, J.A., Spacek, D.V., Snyder, M.P., 2015. High-Throughput Sequencing
  Technologies.
- 802 Roca, F.J., Mulero, I., López-Muñoz, A., Sepulcre, M.P., Renshaw, S.A., Meseguer, J.,
- 803 Mulero, V., 2008. Evolution of the inflammatory response in vertebrates: fish TNF-
- alpha is a powerful activator of endothelial cells but hardly activates phagocytes.
- 805 Journal of immunology (Baltimore, Md. : 1950) 181, 5071-5081.
- 806 Rodriguez-Jorquera, I.A., Colli-Dula, R.C., Kroll, K., Jayasinghe, B.S., Marco, M.V.P.,
- 807 Silva-Sanchez, C., Toor, G.S., Denslow, N.D., 2019. Blood Transcriptomics Analysis of
- 808 Fish Exposed to Perfluoro Alkyls Substances: Assessment of a Non-Lethal Sampling
- Technique for Advancing Aquatic Toxicology Research. Environmental Science &
  Technology 53, 1441-1452.
- 811 Rodriguez-Jorquera, I.A., Kroll, K.J., Toor, G.S., Denslow, N.D., 2015. Transcriptional and
- 812 physiological response of fathead minnows (Pimephales promelas) exposed to urban
- 813 waters entering into wildlife protected areas. Environmental Pollution 199, 155-165.
- 814 Rodriguez-Jorquera, I.A., Silva-Sanchez, C., Strynar, M., Denslow, N.D., Toor, G.S., 2016.
- Footprints of Urban Micro-Pollution in Protected Areas: Investigating the Longitudinal
  Distribution of Perfluoroalkyl Acids in Wildlife Preserves. PLoS One 11.
- 817 Sato, I., Kawamoto, K., Nishikawa, Y., Tsuda, S., Yoshida, M., Yaegashi, K., Saito, N., Liu,
- 818 W., Jin, Y., 2009. Neurotoxicity of perfluorooctane sulfonate (PFOS) in rats and mice
- 819 after single oral exposure. The Journal of toxicological sciences 34, 569-574.
- 820 Scholz, S., Mayer, I., 2008. Molecular biomarkers of endocrine disruption in small
- 821 model fish. Mol Cell Endocrinol 293, 57-70.
- 822 Segner, H., Verburg-van Kemenade, B.M.L., Chadzinska, M., 2017. The
- 823 immunomodulatory role of the hypothalamus-pituitary-gonad axis: Proximate
- mechanism for reproduction-immune trade offs? Developmental & ComparativeImmunology 66, 43-60.
- 826 Shi, X., Yeung, L.W.Y., Lam, P.K.S., Wu, R.S.S., Zhou, B., 2009. Protein Profiles in
- Zebrafish (Danio rerio) Embryos Exposed to Perfluorooctane Sulfonate. ToxicologicalSciences 110, 334-340.
- 829 Shi, X.J., Du, Y.B., Lam, P.K.S., Wu, R.S.S., Zhou, B.S., 2008. Developmental toxicity and
- alteration of gene expression in zebrafish embryos exposed to PFOS. Toxicology and
- Applied Pharmacology 230, 23-32.

- 832 Spulber, S., Kilian, P., Ibrahim, W.N.W., Onishchenko, N., Ulhaq, M., Norrgren, L., Negri,
- 833 S., Di Tuccio, M., Ceccatelli, S., 2014. PFOS induces behavioral alterations, including
- spontaneous hyperactivity that is corrected by dexamfetamine in zebrafish larvae.
  PLoS One 9, e94227-e94227.
- 836 Stegeman, J.J., Goldstone, J.V., Hahn, E., 2010. Perspectives on zebrafish as a model in
- environmental toxicology, in: StF. Perry, M.E., AP Farrell, Colin, J.B. (Eds.), Fish
  Physiology. Academic Press, pp. 367-439.
- 839 Stengel, D., Zindler, F., Braunbeck, T., 2017. An optimized method to assess ototoxic
- effects in the lateral line of zebrafish (Danio rerio) embryos. Comparative Biochemistryand Physiology Part C: Toxicology & Pharmacology 193, 18-29.
- 842 Strahle, U., Scholz, S., Geisler, R., Greiner, P., Hollert, H., Rastegar, S., Schumacher, A.,
- 843 Selderslaghs, I., Weiss, C., Witters, H., Braunbeck, T., 2012. Zebrafish embryos as an 844 alternative to animal experiments--a commentary on the definition of the onset of
- 845 protected life stages in animal welfare regulations. Reprod Toxicol 33, 128-132.
- 846 Stylianou, M., Bjornsdotter, M.K., Olsson, P.E., Jogsten, I.E., Jass, J., 2019. Distinct
- transcriptional response of Caenorhabditis elegans to different exposure routes ofperfluorooctane sulfonic acid. Environmental Research 168, 406-413.
- Sun, G., Liu, K., Wang, X., Liu, X., He, Q., Hsiao, C.D., 2015. Identification and expression
  analysis of zebrafish (Danio rerio) E-selectin during embryonic development. Molecules
  20, 18539-18550.
- Suo, C., Fan, Z., Zhou, L., Qiu, J., 2017. Perfluorooctane sulfonate affects intestinal
  immunity against bacterial infection. Scientific Reports 7, 5166-5166.
- Thienpont, B., Tingaud-Sequeira, A., Prats, E., Barata, C., Babin, P.J., Raldua, D., 2011.
- 855 Zebrafish Eleutheroembryos Provide a Suitable Vertebrate Model for Screening
- 856 Chemicals that Impair Thyroid Hormone Synthesis. Environmental Science &
- 857 Technology 45, 7525-7532.
- van der Aa, L.M., Jouneau, L., Laplantine, E., Bouchez, O., Van Kemenade, L., Boudinot,
- 859 P., 2012. FinTRIMs, fish virus-inducible proteins with E3 ubiquitin ligase activity.
- 860 Developmental & Comparative Immunology 36, 433-441.
- 861 Vasta, G.R., Nita-Lazar, M., Giomarelli, B., Ahmed, H., Du, S., Cammarata, M.,
- 862 Parrinello, N., Bianchet, M.A., Amzel, L.M., 2011. Structural and functional diversity of
- the lectin repertoire in teleost fish: Relevance to innate and adaptive immunity.
- 864 Developmental and Comparative Immunology 35, 1388-1399.
- Vedagiri, U.K., Anderson, R.H., Loso, H.M., Schwach, C.M., 2018. Ambient levels of
  PFOS and PFOA in multiple environmental media. Remediation 28, 9-51.
- 867 Versteeg, G.A., Rajsbaum, R., Sanchez-Aparicio, M.T., Maestre, A.M., Valdiviezo, J., Shi,
- 868 M., Inn, K.S., Fernandez-Sesma, A., Jung, J., Garcia-Sastre, A., 2013. The E3-Ligase TRIM
- Family of Proteins Regulates Signaling Pathways Triggered by Innate Immune Pattern-Recognition Receptors. Immunity 38, 384-398.
- 871 Webster, A.F., Chepelev, N., Gagne, R., Kuo, B., Recio, L., Williams, A., Yauk, C.L., 2015.
- 872 Impact of Genomics Platform and Statistical Filtering on Transcriptional Benchmark
- B73 Doses (BMD) and Multiple Approaches for Selection of Chemical Point of DepartureB74 (PoD). PLoS One 10.
- 875 White, S.S., Fenton, S.E., Hines, E.P., 2011. Endocrine disrupting properties of
- 876 perfluorooctanoic acid. J Steroid Biochem Mol Biol 127, 16-26.
- Woo, J.S., Imm, J.H., Min, C.K., Kim, K.J., Cha, S.S., Oh, B.H., 2006. Structural and
- functional insights into the B30.2/SPRY domain. Embo Journal 25, 1353-1363.

- 879 Yang, L., Allen, B.C., Thomas, R.S., 2007. BMDExpress: a software tool for the
- 880 benchmark dose analyses of genomic data. Bmc Genomics 8.
- Zeng, X.W., Qian, Z., Vaughn, M., Xian, H., Elder, K., Rodemich, E., Bao, J., Jin, Y.H.,
- 882 Dong, G.H., 2015. Human serum levels of perfluorooctane sulfonate (PFOS) and
- 883 perfluorooctanoate (PFOA) in Uyghurs from Sinkiang-Uighur Autonomous Region,
- 884 China: background levels study. Environmental Science and Pollution Research 22,
- 885 4736-4746.
- Zeng, Z.T., Song, B., Xiao, R., Zeng, G.M., Gong, J.L., Chen, M., Xu, P.A., Zhang, P., Shen,
- 887 M.C., Yi, H., 2019. Assessing the human health risks of perfluorooctane sulfonate by in 888 vivo and in vitro studies. Environment International 126, 598-610.
- Zhang, X., Zhao, H., Chen, Y., Luo, H., Yang, P., Yao, B., 2015. A zebrafish (Danio rerio)
- bloodthirsty member 20 with E3 ubiquitin ligase activity involved in immune response
- against bacterial infection. Biochemical and Biophysical Research Communications 457,
- 892 83-89.
- Zhang, Y.H., Wang, J., Dong, G.H., Liu, M.M., Wang, D., Zheng, L., Jin, Y.H., 2013.
- 894 Mechanism of perfluorooctanesulfonate (PFOS)-induced apoptosis in the immunocyte.
- Journal of Immunotoxicology 10, 49-58.

#### 897 *Figure Legends*

898 Figure 1. Morphometric measurements of zebrafish embryos treated with different 899 concentrations of PFOS from 2 to 5 dpf. A) Representative pictures of 900 eleutheroembryos at 5 dpf exposed to different concentrations of PFOS. B) Description 901 of morphological traits measured from lateral (I) and dorsoventral (II) images. 902 Measured traits were the following: body length (BL), head-trunk angle (HTA), yolk sac 903 area (YSA), swim bladder area (SBA), eye length (EL), eye width (EW), head width (HW), 904 inter-ocular distance (IOD) and eye-snout distance (ESD). Scale bar: 1.0 mm. C) 905 Quantitative analyses of the effects of PFOS on the morphological parameters shown 906 in Figure 1C (n=50 larvae/group). All measures are expressed in mm except YSA and SBA (expressed in mm<sup>2</sup>), and HTA (expressed in arc degrees). Different low-case letters 907 908 in each graph indicate statistically differences between experimental groups (Kruskal-909 Wallis tests plus Dunn's pairwise multiple comparisons with Bonferroni correction p < 910 0.05). Means and SEM (standard error of the mean) are represented. Morphological 911 parameters were not determined at the highest concentrations in the exposure (10 912 and 100 mg/L) due to their high rates of mortality (82 and 100 %, respectively). D) 913 Immunofluorescent analyses of olfactory epithelium (A', B') and neuromasts (C', D') 914 structures in 5 dpf eleutheroembryos after exposure to 1 mg/L PFOS. A'-B': Dorsal view 915 of the head of a representative control (A') and 1 mg/L PFOS-exposed (B') 916 eleutheroembryo immunolabeled with a double whole-mount immunofluorescence 917 using anti-parvalbumin (Parv) and anti-acetylated alpha tubulin (a-AT) primary 918 antibodies. The former antibody labels ciliated and microvillous neurons in the sensory 919 olfactory epithelium, whereas the later labels kinocilia of ciliated nonsensory cells and 920 the cilia of the ciliated sensory neurons. The region encircled by a dashed line 921 correspond with the olfactory pits. C'-D': Lateral view (head on the left side) of a 922 representative control (C') and PFOS-exposed (D') eleutheroembryo immunolabeled 923 with anti-acetylated alpha tubulin primary antibody, labeling the neuromasts (nm). 924 Scale bars: 100 µm.

925

Figure 2. Effects of PFOS on zebrafish embryo transcriptome. A) Heatmap showing
concentration changes corresponding to the 1434 transcripts identified by ANOVA-PLS
as differentially expressed genes (DEGs) in at least one of the experimental groups.

929 Values were centered to the average of control samples and log2 transformed. Color 930 scale ranges from blue (strongly underexpressed relative to control) to red (strongly 931 overexpressed); white cells correspond to control values (fold change = 0). Both rows 932 (genes) and columns (samples) were grouped by hierarchical clustering; the 933 corresponding dendrograms are shown at the left and the top of the panel, 934 respectively. B) Results from medoid PAM clustering of DEGs by PCA analysis showing 935 the two defined clusters labeled in blue and red (clusters A and B, respectively). The 936 two first components (PC1 and PC2) of the PCA explained 49.07% of total variability. C) 937 Heatmaps of the genes classified in cluster A (left) and B (right) by the medoid PAM 938 clustering of DEGs; legend as in Figure 2A. D) Normalized abundance values for all the 939 genes included in each of the two clusters (cluster A at left, which contains the 940 underexpressed genes due to PFOS exposure, and cluster B at right, where the genes 941 that enhance their abundance among the exposure are placed). Low-case letters at the 942 top of each graph indicate statistical differences (parametric ANOVA + Tukey's B post-943 hoc test with all pairwise comparisons,  $p \le 0.05$ ). Boxes include values between the 1st 944 and 3rd quartiles, thick bars indicate average values and whiskers cover the total 945 distribution, except for outliers (circles).

946

947 Figure 3. Functional analyses of the genes categorized as DEGs, distributed in clusters 948 as previously shown in Figure 2. A) Distribution of DEGs among the two defined 949 clusters (columns) and the different functional modules (rows). Only clusters with at 950 least four hits in at least one of the clusters are shown. For simplicity, the "Signal" 951 keyword module included in Table 1 was not considered in the figure. Numbers 952 indicate the number of DEGs for each functional module classified in cluster A (left) or 953 B (right). Cell colors represent the relative importance of transcripts associated to each 954 pathway for each cluster (as a heat code: from red -less importance- to white -more 955 importance-). Two cells with the same color correspond to identical fraction of DEGs 956 (i.e. the same importance). Complete David Functional Analysis can be found at Table 957 1. B) Network representation of DEGs, which are represented by dots and colored as 958 before (cluster A in blue and cluster B in red). Networking was carry out according to 959 the adscription of the DEGs to functional modules (GO:biological process and KEGG 960 databases were used, codes for each module are given as nodes). Color ellipses

961 encircle groups of functional modules with particular interest and relevance (see962 section 3.4.).

Category	Term	# of DEGs	Fold Enrichment	p value	FDR
	Cluster 1 (underrepresented in treated sampl	es relative to	controls)		
INTERPRO	IPR001870:B30.2/SPRY domain	24	3.5	4.32E-07	6.66E-04
INTERPRO	IPR006574:SPRY-associated	20	3.4	9.36E-06	0.014
KEGG_PATHWAY	dre04630:Jak-STAT signaling pathway	12	5.2	1.59E-05	0.017
INTERPRO	IPR003877:SPIa/RYanodine receptor SPRY	20	3.2	1.48E-05	0.023
INTERPRO	IPR001507:Zona pellucida domain	9	6.4	7.75E-05	0.120
GOTERM_BP_DIRECT	GO:0016567~protein ubiquitination	16	3.4	9.09E-05	0.133
INTERPRO	IPR003879:Butyrophylin-like	18	3.0	1.10E-04	0.169
GOTERM_BP_DIRECT	GO:0019885~antigen processing and presentation of endogenous	4	25.7	3.52E-04	0.515
INTERPRO	IPR013320:Concanavalin A-like lectin/glucanase	25	2.2	4.27E-04	0.656
GOTERM_MF_DIRECT	GO:0008270~zinc ion binding	47	1.6	0.002	2.160
KEGG_PATHWAY	dre04060:Cytokine-cytokine receptor interaction	11	3.2	0.002	2.341
INTERPRO	IPR011029:Death-like domain	9	3.7	0.003	4.393
GOTERM_BP_DIRECT	GO:0007160~cell-matrix adhesion	6	5.9	0.003	4.807
	Cluster 2 (overrepresented in treated sample	es relative to	controls)		
GOTERM_MF_DIRECT	GO:0008374~O-acyltransferase activity	5	11.5	7.98E-04	1.149
UP_KEYWORDS	Lipid transport	6	7.6	0.001	1.335
GOTERM_BP_DIRECT	GO:0006869~lipid transport	8	5.1	9.50E-04	1.408
KEGG_PATHWAY	dre00564:Glycerophospholipid metabolism	9	4.0	0.002	1.748
	Clusters 1 + 2				
KEGG_PATHWAY	dre04060:Cytokine-cytokine receptor interaction	19	2.6	2.48E-04	0.285
KEGG_PATHWAY	dre04630:Jak-STAT signaling pathway	15	3.1	2.67E-04	0.307
GOTERM_MF_DIRECT	GO:0008374~O-acyltransferase activity	7	7.6	2.13E-04	0.324
UP_KEYWORDS	Signal	218	1.2	9.33E-04	1.178
GOTERM_CC_DIRECT	GO:0005615~extracellular space	40	1.7	0.001	1.531
GOTERM_MF_DIRECT	GO:0004888~transmembrane signaling receptor activity	24	2.1	0.001	2.160
GOTERM_BP_DIRECT	GO:0051607~defense response to virus	8	4.6	0.002	2.354
INTERPRO	IPR011029:Death-like domain	13	2.9	0.002	2.611

**Table 1.** David Functional Analysis<sup>a)</sup> results for both gene clusters, individual and combined (only results with FDR≤5%)

a) https://david.ncifcrf.gov



# Figure 1



A)

Juster A	Cluster B	
0	6	Lipid transport
0	9	dre00564:Glycerophospholipid metabolism
1	8	GO:0006869~lipid transport
2	5	GO:0008374~O-acyltransferase activity
9	0	IPR001507:Zona pellucida domain
4	0	GO:0019885~antigen processing and presentation
6	0	GO:0007160~cell-matrix adhesion
24	0	IPR001870:B30.2/SPRY domain
18	0	IPR003879:Butyrophylin-like
20	0	IPR006574:SPRY-associated
20	0	IPR003877:SPIa/RYanodine receptor SPRY
47	16	GO:0008270~zinc ion binding
5	3	GO:0051607~defense response to virus
9	4	IPR011029:Death-like domain
12	3	dre04630:Jak-STAT signaling pathway
16	4	GO:0016567~protein ubiquitination
25	3	IPR013320:Concanavalin A-like lectin/glucanase
11	8	dre04060:Cytokine-cytokine receptor interaction
15	9	GO:0004888~transmembrane signaling receptor activity



## **Supplementary material**

# Unravelling the mechanisms of PFOS toxicity by combining morphological and transcriptomic analyses in zebrafish embryos

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#### Supplementary methods

#### Supp. method 1: RT-qPCR analysis

Total RNA from 9 replicates (10 larvae per replicate) was extracted using the Trizol method following the manufacturer protocol (Ambion, Thermo Fisher Scientific). Extracted concentrations were assessed with a NanoDrop<sup>®</sup> 8000 UV-Vis Spectrophotometer (Thermo Scientific) and genomic undesirable DNA was removed with Ambion<sup>™</sup> DNase I (Thermo Fisher Scientific). RNA was reverse-transcribed to cDNA using Transcriptor First Strand cDNa Sythesis Kit (Roche Diagnostics) following manufacturer's protocols. A no-RT negative controls were also prepared by replacing the RT enzyme with water. To measure the relative mRNA abundances of target genes, quantitative real-time polymerase chain reactions (qRT-PCR) were carry out, using a LightCycler<sup>®</sup> 480 Real-Time PCR System (Roche Diagnostics, Mannheim, Germany). The amplification program consisted of 10 min at 95 °C, followed by 45 cycles (10 s at 95 °C, 30 s at 60 °C). After the amplification, a dissociation analysis was also programmed (obtaining a melting curve) to evaluate the specificity of the reaction. The amplification reaction efficiency for each gene was assessed (Supplementary Table ST3). Relative transcript abundance were calculated from maximum of the second derivative (Cp, calculated by technical duplicates) of their respective amplification curves. PPIAa was used as a reference gene. Cp values for target genes (Cptg) were normalized to the average Cp values of reference gene, following the equation:  $\Delta Cp_{tg} = Cp_{PPIAa} - Cp_{tg}$ . Transcript abundance changes in samples were calculated by the  $\Delta\Delta$ Cp method [1]:  $\Delta\Delta Cp_{tg} = \Delta Cp_{tg}$  (control group) –  $\Delta Cp_{tg}$  (exposed group). From those  $\Delta\Delta Cptg$  values, foldchange ratios were obtained.

[1] M. W. Pfaffl, "A new mathematical model for relative quantification in real-time RT-PCR," *Nucleic Acids Res.*, 2001.

#### Supplementary Table ST1. RNA-Seq readings and mapping quality statistics.

Sample Barcode	Sample Name	Number of	Million read-pairs	Yield (Gb)	Trimmed	Avg PhiX	Avg PhiX	Organism	Reference	Avg %	Avg %	Avg %	Avg %	Avg %	Avg alignment
Campio Barocao	oumpio numo	FLIs	(or reads)		yield (Gb)	error r1	error r2	erganieni		unique	unmapped	duplicate	difference r1	difference r2	insert size
AB3131	PFOS (2-5) Ctl R1	2	32.569	4.950	4.949	0.25	0.36	Danio rerio	drerio.GRCz10	80.60	2.06	18.38	1.31	1.33	144.00
AB3132	PFOS (2-5) 0,03ppm R1	1	41.359	6.287	6.286	0.17	0.31	Danio rerio	drerio.GRCz10	81.07	1.16	31.63	1.17	1.16	137.00
AB3133	PFOS (2-5) 0,3ppm R1	2	36.463	5.542	5.541	0.16	0.33	Danio rerio	drerio.GRCz10	79.42	1.82	21.92	1.26	1.26	141.00
AB3134	PFOS (2-5) 1ppm R1	2	42.264	6.424	6.422	0.17	0.34	Danio rerio	drerio.GRCz10	82.04	1.45	21.12	1.15	1.17	153.00
AB3139	PFOS (2-5) Ctl R3	2	39.487	6.002	6.000	0.17	0.34	Danio rerio	drerio.GRCz10	81.91	1.64	22.30	1.17	1.18	142.00
AB3140	PFOS (2-5) 0,03ppm R3	2	35.693	5.426	5.424	0.17	0.34	Danio rerio	drerio.GRCz10	83.27	1.52	22.27	1.18	1.20	148.00
AB3141	PFOS (2-5) 0,3ppm R3	1	37.656	5.724	5.717	0.14	0.28	Danio rerio	drerio.GRCz10	82.20	1.44	30.26	1.17	1.29	138.00
AB3142	PFOS (2-5) 1ppm R3	1	48.043	7.303	7.302	0.17	0.31	Danio rerio	drerio.GRCz10	83.20	1.16	30.91	1.24	1.23	137.00
AB3143	PFOS (2-5) Ctl R2	1	36.218	5.505	5.501	NA	NA	Danio rerio	drerio.GRCz10	82.41	1.53	31.48	1.13	1.15	151.00
AB3144	PFOS (2-5) 0,03ppm R2	1	38.790	5.896	5.892	NA	NA	Danio rerio	drerio.GRCz10	81.79	1.25	30.09	1.13	1.15	148.00
AB3145	PFOS (2-5) 0,3ppm R2	1	40.419	6.144	6.139	NA	NA	Danio rerio	drerio.GRCz10	85.39	1.33	32.59	1.16	1.20	153.00
AB3146	PFOS (2-5) 1ppm R2	2	36.776	5.590	5.589	0.16	0.33	Danio rerio	drerio.GRCz10	86.33	1.69	27.59	1.21	1.26	148.50

Column description	
Sample Barcode	Unique internal CNAG identifier of the sample
Sample Name	Sample name provided by the collaborator
Number of FLIs	Number of sequencing Flowcell Lane Index units for this sample (equivalent to the number of times the sample has been sequenced)
Million read-pairs (or reads)	Total million read-pairs (or reads for single-end runs) that passed Illumina filter
Yield (Gb)	Total yield in Gigabases from the read-pairs (or reads for single-end runs) that passed Illumina filter
Trimmed yield (Gb)	Trimmed yield in Gigabases from the read-pairs (or reads for single-end runs) that passed Illumina filter
Avg PhiX error r1	Average % mismatches between the spiked-in PhiX read 1 and the PhiX reference
Avg PhiX error r2	Average % mismatches between the spiked-in PhiX read 2 and the PhiX reference
Organism	Sample's organism according to submitter
Reference	Reference sequence (genome, transcriptome, contigs) to which the reads have been mapped
Avg % unique	Average % of reads aligning to a single location in the reference
Avg % unmapped	Average % of reads not aligning to any location in the reference
Avg % duplicate	Average % of read-pairs (or reads, for single-end runs) that map exactly at the same location in the reference as another read-pair
Avg % difference r1	Average % mismatches between read 1 and the reference
Avg % difference r2	Average % mismatches between read 2 and the reference
Avg alignment insert size	Average mapping distance between read 1 and read 2

**Supplementary Table ST2.** Categorical morphological measurements in 5 dpf eleutheroembryos.

	Sco	liosis	Tail mal- dorsoventra)	formation I and lateral)
Group (mg/L of PFOS)	n/total	Statistical significance <sup>a)</sup>	n/total	Statistical significance
Control	0/50	n/a	4/100	n/a
0.10	0/50	n/a	4/100	n/a
0.25	0/49	n/a	5/98	n/a
0.50	0/50	n/a	6/100	n/a
1.0	0/50	n/a	11/100	n/a
2.5	1/50	n/a	17/100	**
5.0	1/50	**	42/100	***
7.5	18/48	***	45/96	***

a) Fisher exact probability test: p < 0.05 (\*), p < 0.01 (\*\*), p < 0.001 (\*\*\*).

Supplementary Table ST3. Efficiency and accession numbers of qPCR primers used in this study

Gene Name	Accesion nº	Description	Forward-primer	Reverse primer	Efficiency	R <sup>2</sup>
ppiaa	NM_212758.1	peptidylprolyl isomerase Aa (cyclophilin A)	GGGTGGTAATGGAGCTGAGA	AATGGACTTGCCACCAGTTC	92.8%	>0.999
apoa4b.3	XM_001338001.6	apolipoprotein A-IV b, tandem duplicate 3	ACCAAATTAACTCTTCACAAACTTCG	ATGGCATGATAGGATGAATCGAT	92.3%	0.997
btr31	XM_005170606.2	bloodthirsty-related gene family, member 31	CATCATGCCTGCTCCTCACA	GCCGTGTCATGATCCAGAGTC	102.3%	0.982
ccl20b	NM_001113595.1	chemokine (C-C motif) ligand 20b	TGCAATCATTTTCCACACCG	CATAAGACCCGTTCTTGCGTC	98.5%	0.994
erap1a	XM_009305074.3	endoplasmic reticulum aminopeptidase 1a	CTCCGGGCTCCGCAGTAT	AGCAGTGGGACGTTCCTGTC	95.0%	0.996
erap2	NM_001123052.1	endoplasmic reticulum aminopeptidase 2	TCAGACCGGAGATTGCGTG	CCGTTCAGCCTTTTCCAAAC	103.7%	0.994
sps3b	XM_021470242.1	spIA/ryanodine receptor domain and SOCS box containing 3b	ATATGATGGTTGGCATCGGG	CATCTGTGCCCAGCAGACTG	100.2%	0.993
stat1b	NM_200091.2	signal transducer and activator of transcription 1b	AGTGAAAGCTGTCGAGACGGA	CACCGAATGGATCTTGGGTT	103.2%	0.991

Table ST4. David Functional Analysis<sup>a)</sup> results for both gene clusters, individual and combined (only results with FDR≤5%)

Category	Term	Genes	Count	Fold	PValue	FDR
Cluster 1 (down-				Linemient		
INTERPRO	IPR001870:B30.2/SPRY domain	SI:CH211-133H13.1, SPSB1, SI:DKEY-84H14.2, BTR26, BTR29, SI:CH211-247L8.8, SI:DKEY-3H2.3, SI:CH211-120G10.1, ZGC:174180, SI:CH211-76M11.8, TRIM109, SI:DKEY-222H21.8, BTR20, SI:CH211-255G12.8, RSPRY1, ZGC:194906, FTR37, SI:DKEY-61P9.9, FTR53, FTR35, FTR56, SPSB3B, FTR30, BTR31	24	3.5	4.32E-07	6.66E-04
INTERPRO	IPR006574:SPRY-associated	SI:CH211-133H13.1, SI:DKEY-84H14.2, BTR26, BTR29, SI:CH211- 247L8.8, SI:DKEY-3H2.3, SI:CH211-120G10.1, ZGC:174180, SI:CH211-76M11.8, SI:DKEY-222H21.8, BTR20, SI:CH211- 255G12.8, ZGC:194906, FTR37, SI:DKEY-61P9.9, FTR53, FTR35, FTR56, FTR30, BTR31	20	3.4	9.36E-06	0.014
KEGG_PATHWAY	dre04630:Jak-STAT signaling pathway	IRF9, SI:RP71-17116.5, IFNPHI1, STAT1B, IL6ST, IL10RB, IL20RA, LEPA, IL19L, PIK3CD, SOCS3B, IL22RA2	12	5.2	1.59E-05	0.017
INTERPRO	IPR003877:SPla/RYanodine receptor SPRY	SI:CH211-133H13.1, SPSB1, SI:DKEY-84H14.2, BTR26, BTR29, SI:CH211-247L8.8, SI:DKEY-3H2.3, SI:CH211-120G10.1, ZGC:174180, SI:CH211-76M11.8, SI:DKEY-222H21.8, BTR20, SI:CH211-255G12.8, RSPRY1, ZGC:194906, SI:DKEY-61P9.9, FTR53, FTR56, BTR31, SPSB3B	20	3.2	1.48E-05	0.023
INTERPRO	IPR001507:Zona pellucida domain	ZP3C, SI:CH211-39F2.3, ZGC:66449, SI:CH73-181M17.1, SI:DKEY- 239B22.1, SI:CH211-226H7.8, ZGC:153932, SI:CH211-226H7.5, SI:CH211-226H7.6	9	6.4	7.75E-05	0.120
GOTERM_BP_DIRECT	GO:0016567~protein ubiquitination	MRRN1, TRAF1, BTR26, CCNF, HACE1, UBOX5, ASB12A, ASB18, BTR20, UVSSA, KLHL26, CAND2, KLHL38B, FBXO32, SOCS3B, BTR31	16	3.4	9.09E-05	0.133
INTERPRO	IPR003879:Butyrophylin-like	SI:CH211-133H13.1, SI:DKEY-84H14.2, BTR26, BTR29, SI:CH211- 247L8.8, SI:DKEY-3H2.3, SI:CH211-120G10.1, ZGC:174180, SI:CH211-76M11.8, TRIM109, SI:DKEY-222H21.8, BTR20, SI:CH211- 255G12.8, ZGC:194906, SI:DKEY-61P9.9, FTR53, BTR31, FTR56	18	3.0	1.10E-04	0.169
GOTERM_BP_DIRECT	GO:0019885~antigen processing and presentation of endogenous peptide antigen via MHC class I	ERAP1A, ERAP1B, ERAP2, TAPBPL	4	25.7	3.52E-04	0.515
INTERPRO	IPR013320:Concanavalin A-like lectin/glucanase, subgroup	SI:CH211-133H13.1, SPSB1, SI:DKEY-84H14.2, BTR26, BTR29, SI:CH211-247L8.8, SI:DKEY-3H2.3, SI:CH211-120G10.1, ZGC:174180, SI:CH211-76M11.8, TRIM109, SI:DKEY-222H21.8, BTR20, SI:CH211-25G012.8, RSPRY1, ZGC:194906, FTR37, SI:DKEY-61P9.9, FTR53, FTR35, PROS1, FTR56, SPSB3B, FTR30, BTR31	25	2.2	4.27E-04	0.656
GOTERM_MF_DIRECT	GO:0008270~zinc ion binding	MKRN1, TRAF1, ZNFX1, ZRANB1B, ZMAT2, RORC, ISL1L, ADA, FANCL, BTR20, ZFAND5B, RSPRY1, CECR1A, SETMAR, SI:CH211- 244B2.2, CCS, ERAP2, XAF1, ZGC:77880, TRIM63A, FTR79, NPLOC4, THRAA, ERAP1A, ERAP1B, BTR26, BTR29, SI:CH211- 247L8.8, POLE, ESR1, SIRT5, PHF11, CXXC1A, UBOX5, RNF11A, CA4B, CSRP1B, TRIM109, BHMT, FTR37, RNF25, FTR35, FTR53, RNF26, BTR31, FTR30, FTR56	47	1.6	0.002	2.160
KEGG_PATHWAY	dre04060:Cytokine-cytokine receptor interaction	TNFB, IFNPHI1, CCL20B, CCL20A.3, IL6ST, IL10RB, IL20RA, LEPA, CCR6A, TNFRSF18, IL22RA2	11	3.2	0.002	2.341
INTERPRO	IPR011029:Death-like domain	IRAK3, SI:CH211-114L13.9, CARD9, CASP9, SI:DKEY-103E21.5, SI:DKEY-61P9.9, SI:DKEY-29H14.10, SI:CH211-171L17.4, ADGRG1	9	3.7	0.003	4.393
GOTERM_BP_DIRECT	GO:0007160~cell-matrix adhesion	SI:CH211-39F2.3, HPSE, SI:CH73-181M17.1, SI:DKEY-239B22.1, SI:CH73-329N5.6, EPDL1	6	5.9	0.003	4.807
Cluster 2 (up-regulation)	GO:0008374~O-acyltransferase	SOAT2 I CAT GPAT2 YKT6 GPAM	5	11 5	7 98F-04	1 149
UP_KEYWORDS	activity Lipid transport	APOA4B.3, OSBPL2B, APOA4B.2, APOA4B.1, OSBPL10, OSBPL11	6	7.6	0.001	1.335
GOTERM_BP_DIRECT	GO:0006869~lipid transport	APOA4B.3, OSBPL2B, APOA4B.2, APOA4B.1, OSBPL10, APOBB.1, OSBPL11, VTG6	8	5.1	9.50E-04	1.408
KEGG_PATHWAY	dre00564:Glycerophospholipid metabolism	GPD2, PMT, PLA2G12A, LCAT, PTDSS1B, PLA2G12B, GPAT2, GPAT3, GPAM	9	4.0	0.002	1.748

Table ST4 (continued). David Functional Analysis<sup>a)</sup> results for both gene clusters, individual and combined (only results with FDR<5%)

Clusters 1 + 2						
KEGG_PATHWAY	dre04060:Cytokine-cytokine receptor interaction	CCL20B, IL6ST, CCR6A, MET, IL2RGA, TNFB, IL12RB2, IFNPHI1, VEGFC, IL20RA, CCL20A.3, IL10RB, LEPA, TNFRSF18, TNFRSF19, TGFB1A, NGFRB, IL22RA2, THPO	19	2.6	2.48E-04	0.285
KEGG_PATHWAY	dre04630:Jak-STAT signaling pathway	STAT1B, IL6ST, PIK3CD, IL2RGA, IRF9, IL12RB2, SI:RP71-17I16.5, IFNPHI1, IL20RA, IL10RB, LEPA, IL19L, SOCS3B, THPO, IL22RA2	15	3.1	2.67E-04	0.307
GOTERM_MF_DIRECT	GO:0008374~O-acyltransferase activity	SOAT2, LCAT, NAA40, MBOAT4, GPAT2, YKT6, GPAM	7	7.6	2.13E-04	0.324
UP_KEYWORDS	Signal	LTBP1, IL6ST, SI:CH73-181M17.1, LHCGR, PDGFBA, EPDL1, B2M, SI:CH211-222K6.3, OGN, DEFBL1, LCAT, CFH, ZGC:153932, ERAP2, SERPINH1A, LAMB1A, SOSTDC1B, SI:DKEY-4C23.5, CCL34A.3, PKHD1L1, SI:CH211-145B13.6, ALP12, ZGC:100868, VEGFC, NPC2, ZGC:162608, SI:CH211-125E6.11, TGFB1A, ADAMTS5, UGT5G2, ZGC:77929, SI:RP71-1G18.7, IL17C, GNAIA, TMEM123, SI:CH1073-155H21.1, SI:CH1073-126C3.2, ZGC:163079, FREM2A, ZGC:123297, ADGRE14, APOBB.1, CYP2R1, HSPA13, MHC12BA, SI:CH211-134A4.3, SI:DKEY-112E17.1, LXN, NITR1B, LY86, NPNT, ZGC:171509, TAS1R3, SI:DKEY-84K17.2, CCL20A.3, PDFK1A, SCPP8, USP11, PRS355, MBLAC1, SI:CH211-39F2.3, AVP, GIP, JKAMP, CNPY2, P4HA1B, SI:DKEY-162H11.3, CA4B, IL20RA, NITR13, FKBP14, TOR1L1, WNT9A, GPHA2, SI:CH73- 334D15.1, ZGC:172053, HYAL2A, SEMA3GA, TFA, IL10RB, PLA2G12A, PLA2G12B, SEMA3GB, LY75, SI:DKEY-193C22.2, IGF3, STAB2, CLEC11A, ZGC:55621, KHK, CRFB15, FBIA2, SI:CH73- 380L10.2, PCMTL, IGFBP1B, SI:CH211-132F1.4, IL2RGA, DICP2.2, VIPR2, MMP2, LYGL1, PRF1.1, CCL39.6, SI:CH211-125E6.5, SI:CH73-256J6.4, SI:CH211-138H31.1, SPARCL2, FZD7B, SI:DKEY- 192K22.2, EPHB4B, ZGC:111983, SI:DKEY-88N24.8, MFAP5, PLXNC1, IGFBP5B, CTSS2.2, CTSS2.1, ITGB5, SI:CH211-145C1.1, SI:DKEY-247K7.2, IFNPH11, SI:CH73-330K17.3, ANGPTL3, LOXL3A, HAVCR1, GNRH2, MPEG1.2, SI:CH73-329N5.6, SI:DKEY- 239B22.1, MET, ADGRG1, SI:CH27-19G15.3, SI:CH211-76M11.8, SI:DKEY-37G12.1, CXCL19, ADGRF8, ITGA8, SI:DKEY-21E2.8, SLC8A2A, PCOLCE2B, IL22RA2, ANTXR1D, GM2A, HEXA, HDR,	218	1.2	9.33E-04	1.178
GOTERM_CC_DIRECT	GO:0005615~extracellular space	SERPINB1L1, SI:DKEY-23A13.6, CCL20B, IGFBP5B, LUM, CTSS2.2, ZGC:154142, CTSS2.1, ZGC:171509, SEMA3AB, PDGFBA, IGF2A, IL17C, TNFB, OGN, IFNPH11, SEMA3GA, TFA, WNT3, SI:CH211-113A14.19, CCL20A.3, LEPA, CECR1A, C18H30RF33, SEMA3GB, DIA1B, SERPINH1A, SOSTDC1B, SAAL1, INHA, IGF3, VEGFC, SI:CH73-36P18.5, CPAMD8, CTSD, CTSC, SI:CH73-380L10.2, TGFB1A, IGFBP1B, SEMA3FB	40	1.7	0.001	1.531
GOTERM_MF_DIRECT	GO:0004888~transmembrane signaling receptor activity	LY75, OR132-4, ADGRF6, SI:CH211-282J17.8, OR111-11, OR104- 2, FZD7B, VIPR2, ADGRG1, OR103-2, SI:CH211-218M3.18, SI:DKEY-83F18.7, SI:CH211-225K7.3, ADGRE16, ADGRF8, OR105- 1, ADGRE14, SI:CH211-225K7.5, ADGRE5B.2, SI:DKEY-23C22.5, SI:DKEY-88N24.8, PLA2R1, SI:CH211-282J17.1, OR126-3	24	2.1	0.001	2.160
GOTERM_BP_DIRECT	GO:0051607~defense response	MXF, IFNPHI1, BNIP3LA, IFIT14, MXC, SI:CH73-236C18.8, LY86, MXA	8	4.6	0.002	2.354
INTERPRO	IPR011029:Death-like domain	DLG5B.1, SI:CH211-114L13.9, IRAK3, CARD9, CASP9, SI:DKEY- 103E21.5, HDR, SI:DKEY-61P9.9, SI:CH211-66K16.2, SI:DKEY- 29H14 10, SI:CH211-1711 17.4, ADGRG1, NGERB	13	2.9	0.002	2.611
GOTERM_BP_DIRECT	GO:0019885~antigen processing and presentation of endogenous pentide antigen via MHC class I	ERAP1A, ERAP1B, ERAP2, TAPBPL	4	12.8	0.003	4.194

a) https://david.ncifcrf.gov



B)



Supplementary Figure 1. Quantitative analysis of survival, morphometric and transcriptional data. A) Survival, hatching and swim bladder (SB) inflation rates of zebrafish embryos exposed to PFOS in the morphometric test. Measurements were taken at 3, 4 and 5 dpf. Bars represent the mean value ± SEM (standard error of the mean) for each group. Non-parametric test (Kruskal-Wallis with pairwise multiple comparisons, p < 0.05) was performed. Colored asterisks indicated statistical differences between exposed embryos and the corresponding control group for each age. B). Benchmark dose (BMD) analysis for morphometric and transcriptional changes. The graph shows accumulation plots of the best calculated BMD for each gene and morphological trait. BMD was calculated for each parameter by the BMDExpress software (https://www.sciome.com/bmdexpress/), using a combination third-order polynomial, third-order exponential and Hill equations, and choosing the best fitted model for each case separately. In case of the genes, only those with more than 100 counts between all the 12 replicates and a fold change > 1.5 or <0.75 were used. The median values of BMD and BMDL were calculated as 0.027 and 0.011 mg/L for the transcriptomic effects and 2.83 and 2.53 mg/L for the morphological effects, respectively. BMDL values were considered as PoD (Point of Departure) levels, doses at which negative effects started being significant). Dotted lines indicate the range coincidence of transcriptomic and morphological effects (in terms of % of affected genes/traits) between the most and less sensitive parameters.



**Supplementary Figure 2**. Correlation between RNAs-Seq and RT-qPCR relative expression of some selected genes (apoa4b.3, btr31, ccl20b, erap1a, erap2, spsb3b and stat1b). Fold change expression data were used and a linear regression performed. Colored points represent the mean fold change ± SEM (standard error of the mean) of the three treatment groups (0.03, 0.3 and 1.0 mg/L of PFOS) per each gene. Three and nine replicates were used for RNA-Seq and RT-qPCR, respectively.



Fold change expression (PFOS)

**Supplementary Figure 3**. Correlation of the relative expression (measured with RNA-Seq) of the differentially expressed genes (DEGs) related with lipid transport/metabolism between our PFOS exposure and a previous BPA exposure done in the same developmental timing and conditions [33]. For comparison, the most similar groups between exposures were linked with each other (controls between them -not used for the regression-, 1/40 LOAEC of BPA with 1/30 LOAEC of PFOS, 1/4 LOAEC of BPA with 1/3 LOAEC of PFOS and LOAEC of BPA with LOAEC of PFOS). Fold change expression data were used and a linear regression performed. Each point represents the mean value ± SEM (standard error of the mean) of the 3 transcriptomic replicates used in each exposure.







**Supplementary Figure 4**. Analyses of the myosines, tropomyosins and actins categorized as DEGs in the general transcriptomic analyses. A) Heatmap showing concentration changes corresponding to the 12 transcripts (only considering myosines, tropomyosins and actins) identified by ANOVA-PLS as differentially expressed genes (DEGs) in at least

A)

one of the experimental groups. Values were centered to the average of control samples and log2 transformed. The mean value of the 3 replicates is shown for each group. Color scale ranges from blue (strongly underexpressed relative to control) to red (strongly overexpressed); white cells correspond to control values (fold change = 0). Rows (genes) were grouped by hierarchical clustering and its corresponding dendrogram is shown at the top of the panel. The two clusters (A and B) determined in a posterior medoids PAM clustering analysis are indicated. B) Normalized abundance values for all the genes included in each of the two clusters (cluster A at left, which contains the overexpressed genes due to PFOS exposure, and cluster B at right, which contains the underexpressed genes). Low-case letters at the top of each graph indicate statistically different distributions (parametric ANOVA + Tukey's B post-hoc test with all pairwise comparisons,  $p \le 0.05$ ). Boxes include values between the 1st and 3rd quartiles, thick bars indicate average values and whiskers cover the total distribution, except for outliers (circles).